- 1 Immobilizing Laccase on Kaolinite and Its Application in Treatment of Malachite
- 2 Green Effluent with the Coexistence of Cd (II)
- 3 Xiaofeng Wen<sup>a,1</sup>, Chunyan Du<sup>b,c,1</sup>, Jia Wan<sup>a,1</sup>, GuangmingZeng<sup>a\*</sup>, Danlian Huang<sup>a\*</sup>,
- 4 Lingshi Yin<sup>b,c</sup>, Rui Deng<sup>a</sup>, ShiyangTan<sup>b,c</sup>, Jinfan Zhang<sup>b,c</sup>
- <sup>a</sup> College of Environmental Science and Engineering, Hunan University and Key
- 6 Laboratory of Environmental Biology and Pollution Control (Huran University),
- 7 Ministry of Education, Changsha 410082, P.R. China;
- <sup>8</sup> <sup>b</sup> School of Hydraulic Engineering, Changsha University of Science & Technology,
- 9 Changsha 410114, P.R. China;
- <sup>c</sup> Key Laboratory of Water-Sediment ciences and Water Disaster Prevention of Hunan
- 11 Province, Changsha 410114, R. N
- 12

\*Corresponding author.

E-mail address: zgming@hnu.edu.cn (G.M.Zeng), huangdanlian@hnu.edu.cn (D.L.

Huang). Tel.: +86 731 88823701. Fax: +86 731 88823701.

1 These authors contribute equally to this article.

**ABSTRACT:** Malachite green effluent with the Coexistence of Cd (II) was efficiently 13 14 decolorized by kaolinite-laccase (Kaolin-Lac). Laccase from Trametes versicolor was 15 immobilized onto the kaolinite through physical adsorption contact. The optimal conditions were 180 min of immobilization time and 0.8 mg/mL of enzyme solution. 16 17 Kaolin-Lac could obtain a loading efficiency of 88.22%, a loading capacity of 12.25 18 mg/g, and the highest activity of 839.01 U/g. Moreover, the pr s of immobilization increased its pH stability and operational stability. Kaolin 19 ned above 50% of the original activity and nearly 80% decolorizati for MG after 5 cycles. In the 20 presence of 3, 5-Dimethoxy-4-hydroxybe yd (SA), Kaolin-Lac could degrade 21 22 over 98% of malachite green. The coexistence of Cd  $(\Pi)$  was beneficial to the decolorization of malachite Kaolin-Lac. The structural and morphological 23 24 ac and Kaolin-Lac after degradation were determined by features of kaolin scanning electron microscopy-energy spectrum analysis (SEM-EDS) and Fourier 25 26 transform infrared spectroscopy (FTIR). Cadmium appeared on the Kaolin-Lac after 27 degradation. After immobilization and degradation, the surface groups on kaolinite were 28 changed. Kaolin-Lac showed its more potential continuous employment than free 29 laccase in practical malachite green dyes effluent mixed with Cd  $(\Pi)$ .

- **KEYWORDS:** Malachite Green, Cd (II), Coexistence, Laccase immobilization,
- 31 Kaolinite, Removal



## 33 **1. Introduction**

34	Immobilization could improve laccase properties and it is an effective way to
35	overcome the application limitations of laccase, such as low stability and high
36	production costs (Mohamad et al., 2015; Prasad and Palanivelu, 2015). The exploration
37	of immobilization methods of laccase has been predominantly focused (Barbosa et al.,
38	2013; Guzik et al., 2014). The immobilization methods can increase the stability of
39	laccases, thus significantly reducing the cost burden (Datta et al., 2015; Sheldon and van
40	Pelt, 2013).
41	In the laccase immobilization process, from carriers have been reported to
42	immobilize laccase successfully (Ar et a), 2015; Tan et al., 2015). Among various
43	carriers, kaolinite as an alumino-spicate mineral is cost-efficient, facility of reusability,
44	low mass transfermesistance and microbial corrosion resistance (Abdul Rahman et al.,
45	2005; Hu et al., 2007). Kaolinite has negative sites on the basal surface owing to
46	isomorphic substitution and amphoteric sites on the edge surface (Liang et al., 2017;
47	Shu et al., 2016). The amphoteric sites are conditionally charged and pH dependent
48	because a net positive or net negative charge can be produced due to proton adsorption
49	(An et al., 2015; Zhang et al., 2015). Kaolinite has a low permanent charge and a

50	significant variable charge (Sinegani et al., 2005; Xu et al., 2012). Attributing to these
51	distinctive characters, kaolinite has rather high adsorption ability. Kaolinite is widely
52	used in adsorption studies.
53	Malachite green (MG) was produced from the textile staining, aquaculture, food
54	and medical domains (Chen et al., 2015; Sinha and Osborne, 2016). The MG belongs to
55	persistent contaminant, and it can be readily adsorbed on solid consorbed by organism
56	thus leading to the accumulations in organisms (Gong et al., 200). The accumulation of
57	MG hinders organisms' growth, reproduction and development, and they can generate
58	mutagenic and carcinogenic influence. Furthermore, the MG effluent is always released
59	to environment combining with heavy metals like Cd ( $\Pi$ ) in realistic situations (Deng et
60	al., 2013; Jasinska et al., 2012). The Cd (Π) has been listed as one of the top toxic
61	heavy metals since it an ause cancer, bone lesions, lung insufficiency, anemia,
62	hypertension and weight loss (Long et al., 2011; Wan et al., 2018). Elevated level of Cd
63	( $\Pi$ ) could lead to the acute and chronic disorders in nervous, kidney, liver and
64	cardiovascular system, therefore, efficient removal of Cd ( $\Pi$ ) makes sense (Tang et al.,
65	2014; Wu et al., 2017). The mixing of Cd ( $\Pi$ ) makes the MG effluent more difficult to
66	treat (Ren et al., 2018; Xu et al., 2012).

67	The redox-mediated bio-oxidation of MG, catalyzed by immobilized laccase, is a
68	current technology for MG degradation (Kuhar et al., 2015; Zhang et al., 2016).
69	Utilizing laccase immobilized on numerous carriers to catalyze MG degradation is an
70	effective and straight forward method (Zhou et al., 2018). There are varieties of carriers,
71	including chitosan beads, PAN/O-MMT composite nanofibers, a sponge-like hydrogel,
72	and amino-functionalized magnetic nanoparticles were applied to mamobilize laccase to
73	degrade the MG effluent (Kumar et al., 2014; Li et al., 2015; Zheng et
74	al., 2016). However, immobilizing laccase from Transetes versicolor on kaolinite to
75	degrade MG effluent with the coexistence of Cd (II) was not fully explored. The
76	existence of Cd (II) does have effect in the activity of free laccase and the degradation
77	of MG by laccase (Cheng et al., 2007). Furthermore, during the laccase immobilization
78	process, immobilitation arrive also has important effect on laccase activity and stability
79	(Zheng et al., 2016). Hence, whether if utilizing kaolinite to immobilize laccase from
80	Trametes versicolor could degrade MG effectively, if the coexistence of Cd (II) could
81	affect the degradation of MG and what is the remove efficiency of Cd (II) by
82	immobilized laccase on kaolinite should be fully explored.

83 In the current study, kaolinite that has rather high adsorption ability was used for

laccase physical adsorption immobilization. The efficiency of loaded laccase on 84 85 kaolinite was characterized by relative activity (%) and stability studies. 86 Kaolinite-laccase (Kaolin-Lac) was applied in continuous treatment of MG effluent 87 mixed with Cd ( $\Pi$ ) in the presence of redox mediators SA to explore the removal 88 efficiency of MG and Cd (II) by Kaolin- Lac, the effect of Cd (II) on the degradation 89 efficiency of MG by Kaolin- Lac. The detailed characterization hanges of kaolinite, Kaolin-Lac, Kaolin-Lac after treatment were carried out b 90 -EDS. 91 2. Materials and methods 92 2.1. Materials 93 Laccase from Trametes versice 5-Dimethoxy-4-hydroxybenzaldehyde(SA) and 2,2-azino-bis(3-ethylber zothi ne-6-sulfonic acid) (ABTS) were obtained from 94 een (MG) and Kaolinite were provided by Sinopharm 95 Sigma-Aldrich. N Chemical Reagent Shanghai (China). All other chemicals were of analytical grade and 96 97 were used as received without further purification.

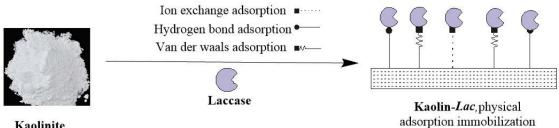
98 **2.2. Enzyme activity assay** 

99 Laccase activity was tested by analyzing the product formation rate of ABTS at the

absorbance of 420 nm. One unit of laccase activity was defined as the amount of free

101 laccase required to oxidize 1  $\mu$ M of substrate per minute. The activity of laccase in this

- 102 study was expressed as the relative activity (%). During the immobilization and stability
- 103 assessing process, the maximal value of laccase activity under each certain condition
- 104 (such as a certain laccase concentration, immobilizing time or pH) was set as 100 %
- 105 (Liu et al., 2012; Zheng et al., 2016).
- 106 **2.3. Immobilization of Laccase**
- 107 The immobilization process was carried out by the physical adsorption contact
- 108 (such as ion exchange adsorption, hydrogen bord adsorption, and Vander waals
- 109 adsorption) of kaolinite and laccase. Kaolinite as lidded in 5 mL of citrate phosphate
- 110 buffer (0.1 M, pH 5) containing lace se (r 01, 0.05, 0.5, 0.8, 1 mg/mL). The mixtures
- 111 were incubated in a rotary shaker at  $2^{\circ}$  C and shake at 200 rpm for 15, 30, 60, 120, 180,
- 112 300 min. Then, the sample was centrifuged and the bottom solid was collected to wash a
- 113 few times with buffer. The final solid was kaolin-*Lac* after freeze drying at -100  $^{\circ}$ C for
- 114 12 h. The Fig. 1 demonstrated the process of immobilization.



115 Kao

116 **Fig.1** The process of laccase immobilized on kaolinite.

#### 117 **2.4. Stability Assessment.**

- 118 2.4.1. pH stability of immobilized laccase
- 119 The pH value has great influence on the laccase activity. For pH stability, the
- 120 samples were added to tubes that contained buffers (pH range of 3 to 6) and incubated at
- 121 30 °C and kept at 200 rpm. The residual enzyme activity of same was determined.
- 122 2.4.2. Thermal stability of immobilized laccase
- 123 The thermal stability of immobilized laccase and free laccase were assessed. Free
- and immobilized laccase were kept from 30, °C 80°C. Then the free and immobilized
- 125 laccase were separately reacted wit ABTS and centrifuged and the absorbance of
- 126 supernatant was measured at +20 r
- 127 2.4.3. Storage stability or impobilized laccase

128 To test the storage stability, the immobilized laccase and free laccase samples were

- 129 stored at 4 °C for 30 days of incubation cycles and residual activities were measured at
- every 5 day.
- 131 2.4.4. Operational stability of immobilized laccase
- 132 Kaolin-Lac was dispersed incitrate-phosphate buffer (pH 5) containing 1 mM

ABTS and cultured for 5 min. The sample was centrifuged and the content oftransformed ABTS in the supernatant was determined. The Kaolin-*Lac* was washed with

- 135 citrate-phosphate buffer, decanted and the procedure was repeated for 5 cycles.
- 136 2.5. Immobilized Laccase System for Treatment of MG Effluent mixed with Cd (II)
- 137 The decolorization efficiency of MG by Kaolin-*Lac* was analyzed by the decrease
- in absorbance at the absorption wavelength of 623 nm. The removal capacity of  $Cd(\Pi)$
- 139 was analyzed by atomic absorption spectrophotometer (AAS). Effect of parameters such
- 140 as the concentration of SA and Cd ( $\Pi$ ), reaction time we studied. The reaction mixture,
- 141 containing kaolin-Lac and 10 mg/L of MG and Ca(1) solution, was incubated at 30 °C.
- 142 After centrifuged, the residual comminuts concentration in the supernatant was
- 143 analyzed. All the experiment, were carmined in triplicate.
- 144 **2.6. Characterization Analytical Methods**

The SEM imaging was applied to characterize kaolinite, kaolin-*Lac* before and after
degradation. The sample was gold-coated using a sputter coater before imaging.
Micrographs were captured at 20 kV accelerating voltage on a scanning electron micro
scope (FEI, QUANTA F250). Elemental analysis was determined with electronic
differential system (EDAX, GENESIS). FT-IR spectra were recorded in the range of

 $500-4000 \text{ cm}^{-1}$  by a Nicole 5700 FT-IR Spectrometer. 150

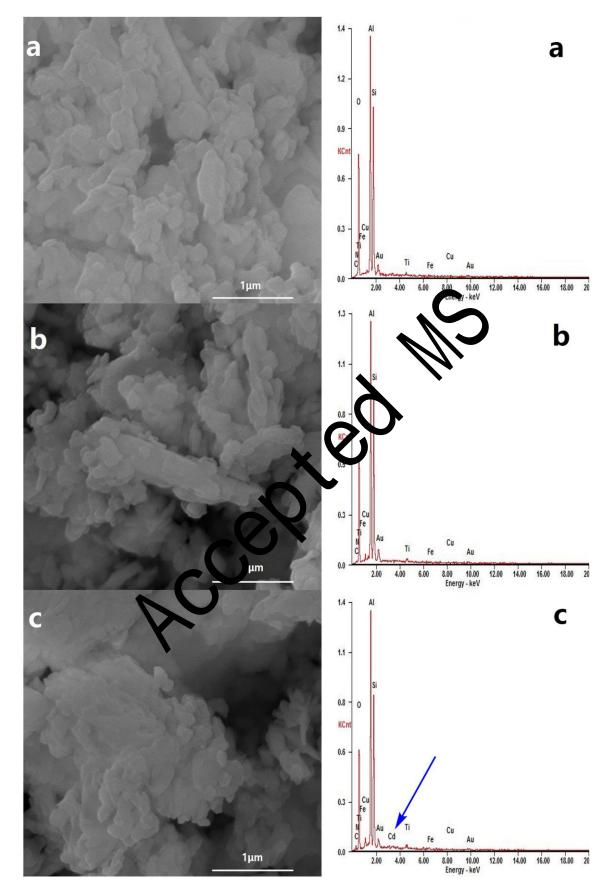
- 151 3. Results and discussion
- 152 3.1. Characterization analysis
- 153 3.1.1. SEM and elements analysis

been displayed d

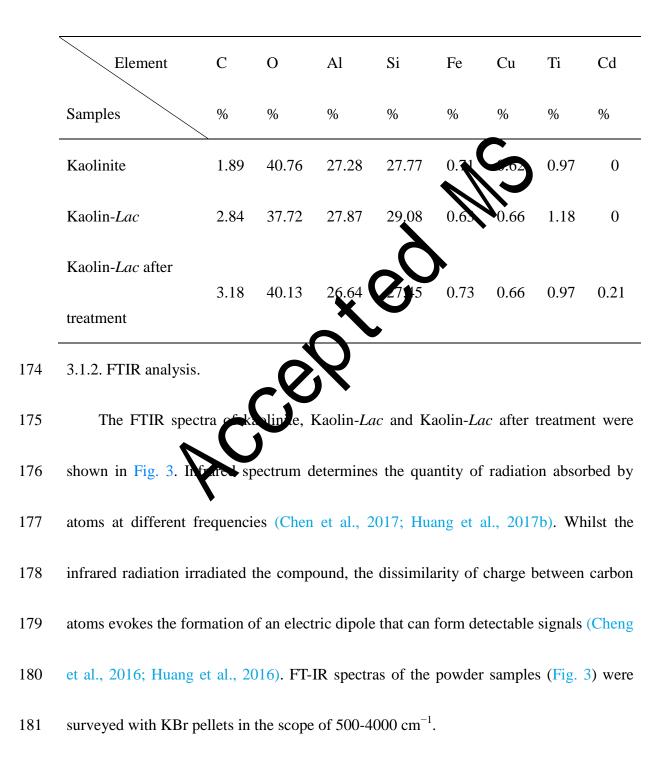
- 154 The SEM aimed to observe possible morphological changes of kaolinite, **N** type swelling clay 155 Kaolin-Lac, and Kaolin-Lac after degradation. Kaolinite, it is a (Huang et al., 2017a; Shu et al., 2016). The kaolinite la 156 are made up of tetrahedral Si-O and octahedral Al-O. They are com d with the van der waals' forces. 157 The structure lamella is filled with commutative 158 ations and water molecules (Huang et 159 al., 2018a; Shu et al., 2014).They n b replaced by ion exchange and inter layer hology of kaolinite and Kaolin-Lac samples has exchange (Shu et al., 2016) 160 ne images (Fig.2). There was no obvious distinction of 161
- different surface morphologies can be observed directly. The relevant EDS confirmed 162
- 163 no obvious elements changed after immobilization in Fig.2a and Fig.2b. However, the
- 164 relevant EDS in Fig.2c revealed elemental difference of Kaolin-Lac after treatment of
- MG effluent with Cd (II). The appearance of cadmium (Cd) in Fig. 2c relevant EDS 165
- visually indicated the Cd ( $\Pi$ ) was adhered to Kaolin-Lac external and internal pore 166

- 167 through adsorption. For the sake of preferably comprehending the formation element of
- 168 materials, the percentages of seven mainly changed elements which are carbon, oxygen,
- 169 sodium, magnesium, aluminum and calcium are listed in Table 1.

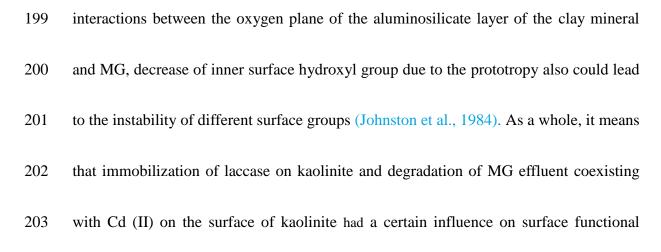


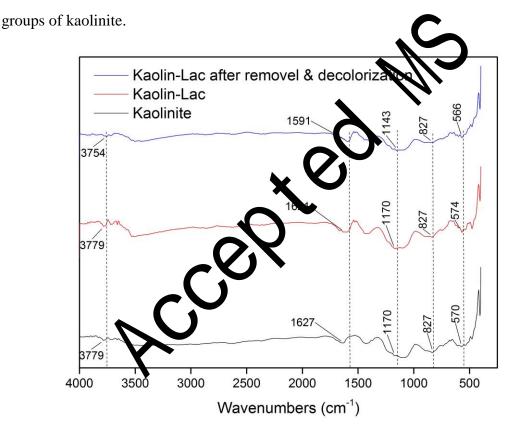


- **Fig. 2.** SEM images and relevant EDS of the a) kaolinite, b) Kaolin-Lac, c) Kaolin-Lac
- after treatment.
- **Table 1**. The element analysis of kaolinite, Kaolin-Lac, Kaolin-Lac after treatment



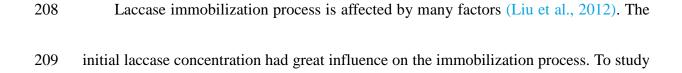
182	Some clear adsorption bands among all samples confirmed that Si-OH and Al-OH
183	were existed in these samples. Among the sample of Kaolinite and Kaolin-Lac, a clear
184	band peak in 3779 cm <sup>-1</sup> can be ascribed to the vibration of Si-OH (Shu et al., 2016; Ztrk
185	et al., 2008). Another adsorption band at 1627 $\text{cm}^{-1}$ belongs to the stretching vibration
186	of H-OH, which was crystal water molecules in the lattice (Wen et al., 2018; Chen et al.,
187	2017; Xue et al., 2018). Absorption bands around 1170 cm <sup><math>-1</math></sup> , $\mathbf{cm}^{-1}$ , $570$ cm <sup><math>-1</math></sup> of
188	spectrum a, b, c are assigned to the bending vibration of Si-transpectively (Huang et al.,
189	2018b; Shu et al., 2016). The adsorption bands among the range of 400-500 cm <sup>-1</sup> were
190	belonging to the group of Si-O and Al-O (Shure a), 2016). After immobilization and
191	treatment, a red shift occurred on the group of Si-O, Si-OH and H-OH. This
192	phenomenon indicated that the survivor groups became unstable after immobilization and
193	treatment (Andjet vi et al. 2015). Due to the "broken-bond" surface, kaolinite has
194	certain of cation exchange capacity. There are a large number of negative charges on
195	Si-O <sup>-</sup> and Al-O <sup>-</sup> groups (Tan et al., 2018). The loading of laccase on kaolinite, the
196	adsorption of Cd (II) on kaolinite and the adsorption or degradation of MG on kaolinite
197	could occupy the binding sites of the exchangeable cations and thus lead to the change
198	on surface charge density of Kaolinite (Huang et al., 2018c). At the same time, the $\pi$





**Fig. 3.** FT-IR spectrum of kaolinite, Kaolin-*Lac*, Kaolin-*Lac* after treatment.

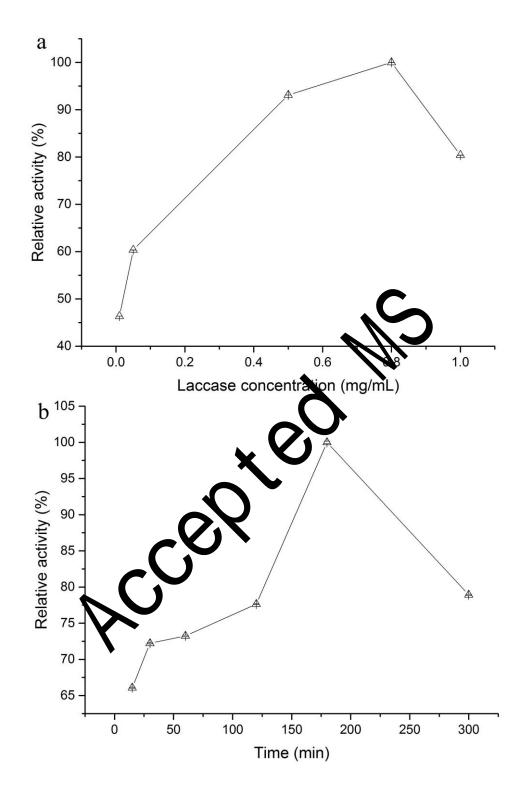
### **3.2. Optimum Conditions of Laccase Immobilization.**



210 the influence of the initial laccase concentration on the activity of immobilized laccase, 211 the initial laccase concentration was varied from 0.01 mg/mL to 1 mg/mL. As shown in 212 Fig.4a, the adsorbed laccase increased with the increase of initial laccase concentration. 213 However, the activity of immobilized laccase increased until 0.8 mg/mL, whilst the 214 laccase concentration was larger than 0.8 mg/mL, a decrease in the relative activity of 215 the Kaolin-Lac was observed. Similar phenomenon was al observed for some previous supporting materials (Kadam et al., 2017; Liu e 216 his could ascribe to the agglomeration or crowding of enzyme mole s onto the surface of kaolinite 217 218 when the laccase was overloaded on the (Oong et al., 2017; Liu et al., 2012). 219 The agglomeration or crowding of accae molecules on surface of kaolinite could constraint the dispersion and trans sion of laccase, even change the conformation of 220 change of laccse activity. Thus, the appropriate laccase 221 laccse and thus concentration was set as 0.8 mg/mL for next studies. 222 223 The immobilization time also remarkably influences the immobilization laccase 224 activity (Huang et al., 2017c; Zheng et al., 2016). As demonstrated in Fig.4b, the 225 relative activity of Kaolin-Lac changed with the immobilization time increased from 15

226 min to 300 min. The relative activity of Kaolin-Lac increased strikingly until 180min,

227 and then the relative activity decreased (Huang et al., 2018d). Activity of immobilized 228 enzymes seemed to rely on the nature of the protein. With time increasing, the possible 229 inactivation amounts of laccase increased during immobilization. In the meantime, the 230 laccase flexibility also reduced. As physical adsorption immobilization time increased, 231 the adsorption site on kaolinite was eliminated (Huang et al., 2017d). The relevant steric 232 hindrance and diffusion limitations might also lead to loss of la ease activity (Prasad and Palanivelu, 2015). The optimum immobilization time 233 nin. Under above optimal condition, the acquired immobilized lace (Kaolin-Lac) could achieve a 234 of  $\mathbf{\hat{p}}$ .55 mg/g, and the highest enzyme loading efficiency of 88.22%, a loading capacit 235 236 activity of 839.01 U/g. ~CCE



237

Fig.4. a) Effect of laccase concentration on the activity of the immobilized laccase; b)

239 Effect of time on the activity of the immobilized laccase.

#### 240 **3.3. Properties of Immobilized Laccase.**

241 3.3.1. pH stability.

242 The pH stability of immobilized laccase is a key factor for practical use (Mahmoodi et al., 2014). The results presented in Fig.5a showed that the Kaolin-Lac 243 244 exhibited more stable relative activity over a broad pH range from 3.0 to 6.0. The minimum relative activity of Kaolin-Lac could maintain mo han 60%. And the 245 relative activity of free laccase was strikingly decreased f 246 When pH value was 6, the relative activity of free laccase almost eased to 2%. The phenomenon 247 suggested that immobilization of laccase 248 vec its pH stability and enhanced its 249 adaptive capacity to the environment The change of pH stability could be ascribed to the proton production on the surfa support. Relative stable proton production could 250 The laminated structure of kaolinite maintained a balance 251 retain the activity proton quantity in the microenvironment around the laccase. This causation could 252 253 explain the higher pH stability of Kaolin-Lac (Skoronski et al., 2017).

- 254 3.3.2. Thermal stability.
- The thermal stability of immobilized laccase is important for its practical application as biocatalyst (Li et al., 2016). As showed in Fig.5b, when the temperature

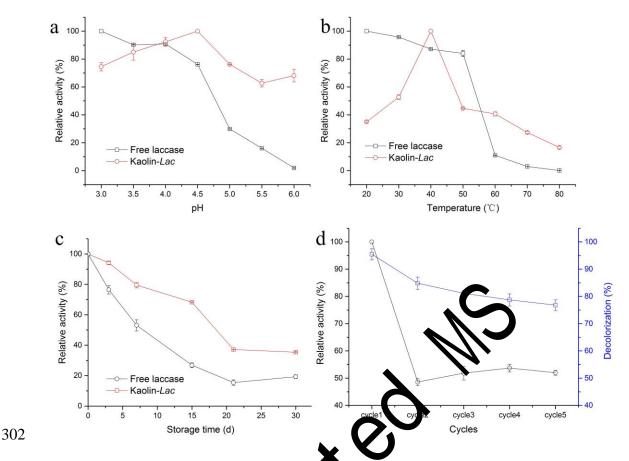
257	was lower than 40 $$ °C, the activity of Kaolin-Lac was increasing with the increase of
258	temperature. Kaolin-Lac showed its highest stability at 40 $$ °C. Then the activity of
259	Kaolin-Lac began to decrease. From 20 $^{\circ}$ C to 80 $^{\circ}$ C, the relative activity of free laccase
260	was decreased all the way. Especially, the activity of free laccase strikingly decreased
261	when temperature was higher than 50 $^{\circ}$ C. And free laccase nearly lost its activity when
262	the temperature was higher than 70 °C. However, when temperature was higher than
263	50 °C, the Kaolin-Lac decreased more tardily and Kaolin-Lac could still maintain 18%
264	of its original activity when temperature was 80 °C eviously, immobilization process
265	improved the tolerance of laccase for high temperatule. The phenomenon was attributed
266	to the augment of enzyme rigidity and decrease of conformational enzyme flexibility
267	which was caused by the righ scality towards denaturation on enzyme by high
268	temperatures. Similarly, the results can be ascribed to the stronger physical bond
269	between the supports and enzyme or the lower restriction of substrate diffusion when
270	the temperature was high (Andjelkovi et al., 2015). The improved thermal stability of
271	immobilized laccase is beneficial to its practical application because the industrial waste
272	water was usually high temperatures (Chang et al., 2016).

273 3.3.3. Storage stability

274	Storage stability is of key influence on the practical utilize of immobilized enzyme
275	(Wang et al., 2013). Generally, the free enzyme is not stable and it could lose its activity
276	by degrees during storage period (Chang et al., 2016). The free and immobilized laccase
277	were kept at 4 for 30 days and their activities were measured per five days to analyze
278	their storage stability. The results were depicted in Fig. 5c and confirmed that the
279	Kaolin-Lac had higher storage stability than the free laccase Daring the whole 30 days,
280	nearly 40% activity was retained for Kaolin-Lac whilst almost only 20% activity of free
281	laccase was retained. Similar increases of storagestability have been reported by Asgher
282	et al. and Ghiaci et al. after immebilizing laccase on chitosan microspheres and
283	bentonite (Asgher et al., 2017, Ghasilet al., 2009). The improvement in storage stability
284	of immobilized locase can be ascribed to the structural rigidness of laccase, the
285	stabilization of support and the protection of laccase from unfolding and denaturation
286	by supports (Andjelkovi et al., 2015).

- 287 3.3.4. Operational stability and reusability
- 288 The operational stability and reusability of immobilized laccase are of important
- 289 significances for practical use. The operational stability and reusability of Kaolin-Lac

290	was explored through a series of cyclic experiments using ABTS as standard substrate
291	and degraded MG effluent with the coexistence of Cd (II). The results shown in Fig. 5d,
292	the Kaolin-Lac lost nearly 50% of its activity after 2 cycles and then the Kaolin-Lac
293	could maintain nearly 50% of its initial activity. The Kaolin-Lac could retain nearly 80%
294	decolorization for MG after five cycles. The physical adsorption immobilization had
295	weak bonds between enzyme and support, hence, the loss of mayme activity can be
296	ascribed to the enzyme leaching during washing process (Linet 11, 2012). However, the
297	laccase sited in the slit of kaolinite was difficult to teck, which resulted in a relatively
298	stable operational stability (Zheng et al., 2016, The stable operational stability could
299	afford stable degradation ability for 1G. immobilization could provide protection for
300	laccase through the carrier haoline and retained laccase activity and the degradation
301	ability for MG.



**Fig. 5.** a) pH stability of free laccase and Kao in-*Lac*; b) Thermal stability studies of free

- 304 laccase and Kaolin-*Lac* at 20-80 Cvp to 20 min; c) Storage stability of free laccase and
- Kaolin-*Lac*; d) Operational stability of Kaolin-*Lac* and reusability of Kaolin-*Lac* for MG
  with the coexistence of Ci (II) decolorization.
- 307 **3.4. Treatment of MG Effluent mixing with Cd (Π).**
- 308 3.4.1. Effect of concentration of SA
- 309 The existence of mediator SA could markedly improve the react velocity. The
- 310 mediator could decrease the energy required for the reaction need, thus accelerating
- 311 reaction (Murugesan et al., 2009). The Fig. 6a and Fig. 6b demonstrated the effect of

312 different concentrations of mediator SA on the decolorization of MG mixed with Cd  $(\Pi)$ . 313 As shown in the Fig.6a, the decolorization of MG had been strikingly improved while 314 the concentration of SA changed from 0 to 0.5 mg/L. However, the decolorization of MG by Kaolin-Lac began to decrease when the concentration of SA exceeded 0.5 mg/L. 315 316 This phenomenon demonstrated that the concentration of SA was important for 317 decolorization of MG by Kaolin-Lac. And the optimum con Relation of SA could provide moderate amount for the oxidation reduction 318 ccase which was immobilized on the kaolinite (Chhabra et al., 2009 uperfluous mediator, SA, which 319 320 couldn't be oxidized by laccase might passes ity to laccase. In the Fig. 6b, the 321 concentration of SA had little influence of the removal of Cd ( $\Pi$ ). This result can be  $(\Pi)$  was the adsorption function of kaolinite and ascribed to which the remov 322 ty had nothing with the activity of laccase, hence, the 323 laccase. The adso concentration of SA had little effect on the removal of Cd (II). 324

325 3.4.2. Effect of concentration of Cd  $(\Pi)$ 

326 Generally, metal ions could bind with the enzyme and change the activity of

- 327 enzyme (Jadhav et al., 2010). Different types and concentrations of metal ions have
- 328 different influence on the laccase activity. Most of them had negative effect on laccase.

329	But several studies confirmed that free laccase had tolerance to different concentration
330	of Ca ( $\Pi$ ), Mg ( $\Pi$ ), Cu ( $\Pi$ ) and Mn ( $\Pi$ ) (Casas et al., 2009; Wang et al., 2017). As a
331	common deleterious metal, Cd ( $\Pi$ ) is often mixed in the textile industries. Some studies
332	have reported that free laccase has tolerance to Cd ( $\Pi$ ) (Casas et al., 2009; Jadhav et al.,
333	2010). However, there were few reports about the influence of Cd ( $\Pi$ ) on the
334	degradation of MG effluent by immobilized laccase. The property of laccase may
335	change after immobilization. The effects of Cd (II) on activity of laccase maybe also
336	change. As depicted in the Fig. 6c, when the concentration of Cd ( $\Pi$ ) was lower than 5
337	mg/L, the decolorization of Kaolin-Lac for MC increased as concentration of Cd $(\Pi)$
338	increased. When the concentration $Ca(\Pi)$ ranged from 5 mg/L to 20 mg/L, the
339	decolorization of Kaolin-Lactor NC was unstable. When the concentration of Cd $(\Pi)$
340	was higher than 12 mg/, the decolorization of Kaolin-Lac for MG maintained with
341	nearly 75%. This phenomenon demonstrated that low contration of Cd (II) could
342	improve the decolorization of Kaolin-Lac for MG, and high concentration of Cd (II)
343	didn't strikingly decrease the decolorization of Kaolin-Lac for MG. It indicated that
344	Kaolin-Lac also had certain tolerance to Cd ( $\Pi$ ) and the immobilization process had
345	little effect on the tolerance of laccase for Cd ( $\Pi$ ). From the Fig. 6d, it could be seen that

Kaolin-Lac had certain removal ability for Cd ( $\Pi$ ). Comparing with kaolinite, the 346 347 Kaolin-Lac had higher removal ability for Cd ( $\Pi$ ). When the concentration of Cd ( $\Pi$ ) 348 was 40 mg/L, Kaolin-Lac had the highest removal efficiency of Cd ( $\Pi$ ), which was 349 close to 23%. The higher removal ability for Cd ( $\Pi$ ) of Kaolin-*Lac* could be attributed 350 to the synergetic effect of kaolinite and laccase (Yang et al., 2017). Immobilization of 351 laccase on kaolinite might change the surface charge density, sy **Tace** functional groups of kaolinite that confirmed by FTIR results, leading to the 352 adsorption ability of kaolinite. Also, laccase could interact with (II) through the Vander waals 353 354 adsorption or hydrogen bond adsorption. 355 3.4.3. Effect of time The Fig. 6e and Fig. 6 the decolorization of MG mixing with Cd  $(\Pi)$  in 356 By contrast, the decolorization of MG and the removal of 357 a 5 h successive e Cd ( $\Pi$ ) by kaolinite were compared. As shown in Fig. 6e, the decolorization of MG by 358 359 Kaolin-Lac could achieve 74% after 15 min incubation. As the decolorization time 360 increased, the decolorization velocity became slow. After 300 min incubation, MG was nearly decolorized totally. However, the kaolinite only achieved 10% decolorization 361 362 after 5 h reaction. This phenomenon could confirm the more important contribution of 363 the laccase catalytic process for decolorization of MG. And with time increasing, the lower decolorization velocity might attribute to the effect of the accumulation of 364 365 degradation products on laccase activity. As depicted in Fig. 6f, the removal of Cd  $(\Pi)$ 366 by Kaolin-Lac and kaolinite demonstrated 21% and 19%, separately. The higher 367 removal ability of Kaolin-Lac might be ascribed to the adsorption of ions by laccase and the change of adsorption ability of kaolinite (Chen et al., 2017 368 ng et al., 2017). As shown in Fig. 6, the removal of MG and Cd  $(\Pi)$  could be a 369 o the united effects of degradation by the laccase and the adsorption by kaolinite support (Yang et al., 370 371 2017).



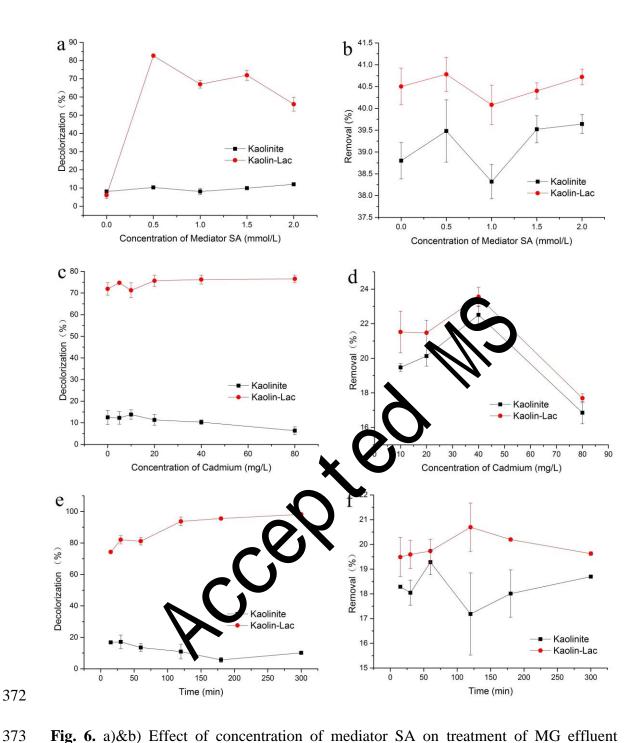


Fig. 6. a)&b) Effect of concentration of mediator SA on treatment of MG effluent
mixing with Cd (Π) by Kaolin-*Lac*; c)&d)Effect of concentration of cadmium on
treatment of MG effluent mixing with Cd (Π) by Kaolin-*Lac*; e)&f)Effect of time on
removal rates of MG effluent mixing with Cd (Π) by Kaolin-*Lac*.

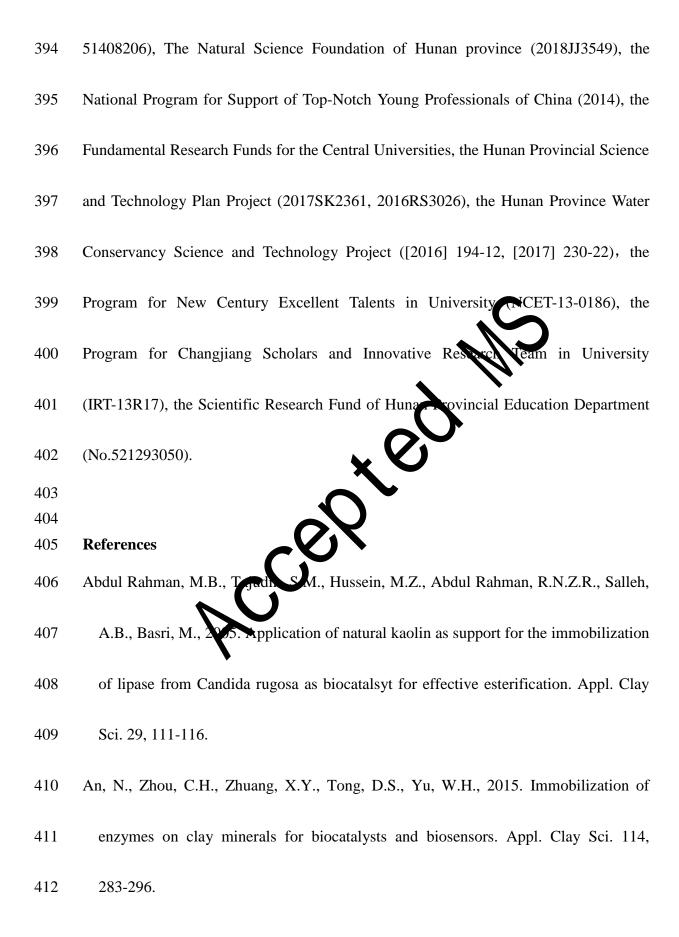
## 377 **4.** Conclusion

378	This study demonstrated the gradual progress of immobilizing laccase on kaolinite.
379	Successful preparation of Kaolin-Lac was confirmed by structural characterizations
380	using FT-IR, SEM-EDS. The stabilities of Kaolin-Lac were enhanced compared to free
381	laccase. The Kaolin-Lac retained 50% of the original activity and nearly 80%
382	decolorization for MG after 5 cycles. In the presence of media SA, the Kaolin-Lac
383	used to degrade MG exhibited nearly 100% in 300 minute, and almost 25% removal
384	for Cd ( $\Pi$ ). This study confirms that kaolinite on serve as microreactor for
385	biomacromolecule immobilization. The low concentration coexistence of Cd (II) could
386	improve the degradation of MG by Kaohn-Lac. Kaolin-Lac had certain tolerance to
387	high concentration of Cd ( $\Pi$ . As chained Kaolin-Lac is an economically, ecofriendly
388	biocatalyst, it has extensive applicability for the treatment of MG effluent mixing with
389	Cd (II).

390

# 391 Acknowledgements

This study was financially supported by the Program for the National Natural
Science Foundation of China (51521006, 51579098, 5110916, 51378190, 51278176,



- 413 Andjelkovi, U., Milutinovi Nikoli, A., Jovi Jovi I, N.A., Bankovi, P., Bajt, T., Mojovi,
- 414 Z., Vuj I, Z., Jovanovi, D.A., 2015. Efficient stabilization of Saccharomyces
- 415 cerevisiae external invertase by immobilisation on modified beidellite nanoclays.
- 416 Food Chem. 168, 262-269.
- 417 Asgher, M., Noreen, S., Bilal, M., 2017. Enhancing catalytic functionality of Trametes
- 418 versicolor IBL-04 laccase by immobilization on chitosan microspheres. Chem. Eng.
- 419 Res. Des. 119, 1-11.
- 420 Barbosa, O., Torres, R., Ortiz, C., Berenguer Murcia, N., Rodrigues, R.C.,
- 421 Fernandez-Lafuente, R., 2013. Heterofunctional Supports in Enzyme Immobilization:
- 422 From Traditional Immobilization Prot cols to Opportunities in Tuning Enzyme
- 423 Properties. Biomacromole ules. 4,2433-2462.
- 424 Casas, N., Parella T., Nicen, T., Caminal, G., Sarra, M., 2009. Metabolites from the
- 425 biodegradation of triphenylmethane dyes by Trametes versicolor or laccase.
- 426 Chemosphere. 75, 1344-1349.
- 427 Chang, Y., Lee, J., Liu, K., Liao, Y., Yang, V., 2016. Immobilization of fungal laccase
- 428 onto a nonionic surfactant-modified clay material: application to PAH degradation.
- 429 Environ. Sci. Pollut. R. 23, 4024-4035.

- 430 Chen, J., Leng, J., Yang, X., Liao, L., Liu, L., Xiao, A., 2017. Enhanced Performance of
- 431 Magnetic Graphene Oxide-Immobilized Laccase and Its Application for the
- 432 Decolorization of Dyes. Molecules. 22, 221
- 433 Chen, M., Xu, P., Zeng, G., Yang, C., Huang, D., Zhang, J., 2015. Bioremediation of
- 434 soils contaminated with polycyclic aromatic hydrocarbons, petroleum, pesticides,
- 435 chlorophenols and heavy metals by composting: Application, nicrobes and future
- 436 research needs. Biotechnol. Adv. 33, 745-755.
- 437 Chen, Y., Peng, J., Xiao, H., Peng, H., Bu, L., Pan, E., He, Y., Chen, F., Wang, X., Li,
- 438 S., 2017. Adsorption behavior of hydrotalcit alke modified bentonite for Pb<sup>2+</sup>, Cu<sup>2+</sup>
- 439 and methyl orange removal from wter. ppl. Surf. Sci. 420, 773-781.
- 440 Cheng, M., Zeng, G., Huan, D., E, C., Xu, P., Zhang, C., Liu, Y., 2016. Hydroxyl
- 441 radicals based advalged oxidation processes (AOPs) for remediation of soils
- 442 contaminated with organic compounds: A review. Chem. Eng. J. 284, 582-98.
- 443 Cheng, Y., He, H., Yang, C., Zeng, G., Li, X., Chen, H., Yu, G., 2016. Challenges and
- 444 solutions for biofiltration of hydrophobic volatile organic compounds. Biotechnol.
- 445 Adv. 34, 1091-1102.
- 446 Chhabra, M., Mishra, S., Sreekrishnan, T.R., 2009. Laccase/mediator assisted

- 447 degradation of triarylmethane dyes in a continuous membrane reactor. J. Biotechnol.
- 448 143, 69-78.
- 449 Datta, S., Christena, L.R., Rajaram, Y.R.S., 2013. Enzyme immobilization: an overview
- 450 on techniques and support materials. 3 Biotech. 3, 1-9.
- 451 Deng, J., Zhang, X., Zeng, G., Gong, J., Niu, Q., Liang, J., 2013. Simultaneous removal
- 452 of Cd(II) and ionic dyes from aqueous solution using mentic graphene oxide
- 453 nanocomposite as an adsorbent. Chem. Eng. J. 226, 189-29
- 454 Ghiaci, M., Aghaei, H., Soleimanian, S., Sedaghat, 2009. Enzyme immobilization:
- 455 Part 2. Immobilization of alkaline phospedas on Na-bentonite and modified
- 456 bentonite. Appl. Clay Sci. 43, 308 16.
- 457 Gong, J.L., Wang, B., Zeng, G.M., Yang, C.P., Niu, C.G., Niu, Q.Y., Zhou, W.J., Liang,
- 458 Y., 2009. Removel of vatichic dyes from aqueous solution using magnetic multi-wall
- 459 carbon nanotube nanocomposite as adsorbent. J. Hazard. Mater. 164, 1517-22.
- 460 Gong, X., Huang, D., Liu, Y., Zeng, G., Wang, R., Wan, J., Zhang, C., Cheng, M., Qin,
- 461 X., Xue, W., 2017. Stabilized Nanoscale Zerovalent Iron Mediated Cadmium
- 462 Accumulation and Oxidative Damage of *Boehmeria nivea (L.) Gaudich* Cultivated in
- 463 Cadmium Contaminated Sediments. Environ. Sci. Technol. 51, 11308-11316.

- 464 Guzik, U., Hupert-Kocurek, K., Wojcieszyska, D., 2014. Immobilization as a Strategy
- 465 for Improving Enzyme Properties-Application to Oxidoreductases. Molecules. 19,
  466 8995-9018.
- 467 Hu, X., Zhao, X., Min Hwang, H., 2007. Comparative study of immobilized Trametes
- 468 versicolor laccase on nanoparticles and kaolinite. Chemosphere. 66, 1618-1626.
- 469 Huang, D., Deng, R., Wan, J., Zeng, G., Xue, W., Wen, X., Zhan, C., Hu, L., Liu, X.,
- 470 Xu, P., Guo, X., Ren, X., 2018. Remediation of lead contaminated sediment by
- 471 biochar-supported nano-chlorapatite: Accompanies with the change of available
- 472 phosphorus and organic matters. J. Hazard. Mater. B48, 109-116.
- 473 Huang, D., Gong, X., Liu, Y., Zeng, L., La, C., Bashir, H., Zhou, L., Wang, D., Xu, P.,
- 474 Cheng, M., Wan, J., 2017 Effect of calcium at toxic concentrations of cadmium in
  475 plants. Planta. 205, 808-871.
- 476 Huang, D., Guo, X., Peng, Z., Zeng, G., Xu, P., Gong, X., Deng, R., Xue, W., Wang, R.,
- 477 Yi, H., Liu, C., 2018. White rot fungi and advanced combined biotechnology with
- 478 nanomaterials:promising tools for endocrine-disrupting compounds biotransformation.
- 479 Crit. Rev. Biotechnol. 38, 671-689.
- 480 Huang, D., Hu, C., Zeng, G., Cheng, M., Xu, P., Gong, X., Wang, R., Xue, W., 2017.

- 481 Combination of Fenton processes and biotreatment for wastewater treatment and soil
- 482 remediation. Sci. Total Environ. 574, 1599-1610.
- 483 Huang, D., Hu, Z., Peng, Z., Zeng, G., Chen, G., Zhang, C., Cheng, M., Wan, J., Wang,
- 484 X., Qin, X., 2018. Cadmium immobilization in river sediment using stabilized
- 485 nanoscale zero-valent iron with enhanced transport by polysaccharide coating. J.
- 486 Environ. Manage. 210, 191-200.
- 487 Huang, D., Liu, L., Zeng, G., Xu, P., Huang, C., Deng, D. Winx, K., Wan, J., 2017.
- 488 The effects of rice straw biochar on indigenous provide community and enzymes
- 489 activity in heavy metal-contaminated sedimeter Chemosphere. 174, 545-553.
- 490 Huang, D., Wang, R., Wang, X., Zhang, C. Zeng, G., Peng, Z., Zhou, J., Cheng, M., Hu,
- 491 Z., Qin, X., 2017. Sorptive receival of ionizable antibiotic sulfamethazine from
- 492 aqueous solution by graphene oxide-coated biochar nanocomposites: Influencing
- 493 factors and mechanism. Chemosphere. 186, 414-421.
- 494 Huang, D., Xue, W., Zeng, G., Wan, J., Chen, G., Huang, C., Zhang, C., Cheng, M., Xu,
- 495 P., 2016. Immobilization of Cd in river sediments by sodium alginate modified
- 496 nanoscale zero-valent iron: Impact on enzyme activities and microbial community
- 497 diversity. Water Res. 106, 15-25.

498	Huang, D.	, Yan, X	K., Yan	. M.	, Zeng,	G.,	Zhou,	C.,	Wan,	J.,	Cheng.	М.,	Xue.	W.,	2018.
		,, _	,	,	,,	<i>–</i> .,	,	<i>–</i> .,	, , , , , , , , , , , , , , , , , , , ,	• • •	B,	,	,,	••••	

- 499 Graphitic Carbon Nitride-Based Heterojunction Photoactive Nanocomposites:
- 500 Applications and Mechanism Insight. ACS Appl. Mater. Inter. 10, 21035-21055.
- 501 Jadhav, J.P., Kallyani, D.C., Telke, A.A., Phugare, S.S., Govindwar, S.P., 2010.
- 502 Evaluation of the efficacy of a bacterial consortium for the removal of color,
- 503 reduction of heavy metals, and toxicity from textile dy Sfluent. Bioresource
- 504 Technol. 101, 165-173.
- Jasinska, A., Rozalska, S., Bernat, P., Paraszkiewicz, Dlugonski, J., 2012. Malachite
- 506 green decolorization by non-basidiomycer framentous fungi of Penicillium
- 507 pinophilum and Myrothecium roridum. ht. Biodeter. Biodegr. 73, 33-40.
- Johnston, C.T., Sposito, G., Jocian, J.F., Birge, R.R., 1984. Vibrational spectroscopic
- study of the interlanellar kaolinite-dimethyl sulfoxide complex. The Journal of
  Physical Chemistry 88, 5959-5964.
- 511 Kadam, A.A., Jang, J., Lee, D.S., 2017. Supermagnetically Tuned Halloysite Nanotubes
- 512 Functionalized with Aminosilane for Covalent Laccase Immobilization. ACS Appl.
- 513 Mater. Inter. 9, 15492-15501.
- 514 Kuhar, F., Castiglia, V., Levin, L., 2015. Enhancement of laccase production and

- 515 malachite green decolorization by co-culturing Ganoderma lucidum and Trametes
- 516 versicolor in solid-state fermentation. Int. Biodeter. Biodegr. 104, 238-243.
- 517 Kumar, V.V., Sivanesan, S., Cabana, H., 2014. Magnetic cross-linked laccase
- 518 aggregates Bioremediation tool for decolorization of distinct classes of recalcitrant
- 519 dyes. Sci. Total Environ. 487, 830-839.
- 520 Li, G.H., Nandgaonkar, A.G., Lu, K.Y., Krause, W.E., Lucia, A., Wei, Q.F., 2016.
- 521 Laccase immobilized on PAN/O-MMT composite nanotibels support for substrate
- 522 bioremediation: a de novo adsorption and biochalytic synergy. RSC Adv. 6,
- 523 41420-41427.
- 524 Liang, J., Yang, Z., Tang, L., Zeng, K., Y., M., Li, X., Wu, H., Qian, Y., Li, X., Luo,
- Y., 2017. Changes in heavy head mobility and availability from contaminated
  wetland soil mediated with combined biochar-compost. Chemosphere. 181,
- 527 281-288.
- 528 Liu, Y.Y., Zeng, Z.T., Zeng, G.M., Tang, L., Pang, Y., Li, Z., Liu, C., Lei, X.X., Wu,
- 529 M.S., Ren, P.Y., Liu, Z.F., Chen, M., Xie, G.X., 2012. Immobilization of laccase on
- 530 magnetic bimodal mesoporous carbon and the application in the removal of phenolic
- 531 compounds. Bioresource Technol. 115, 21-26.

- 532 Long, F., Gong, J., Zeng, G., Chen, L., Wang, X., Deng, J., Niu, Q., Zhang, H., Zhang,
- 533 X., 2011. Removal of phosphate from aqueous solution by magnetic Fe-Zr binary
- 534 oxide. Chem. Eng. J. 171, 448-455.
- 535 Mahmoodi, N.M., Arabloo, M., Abdi, J., 2014. Laccase immobilized manganese ferrite
- 536 nanoparticle: Synthesis and LSSVM intelligent modeling of decolorization. Water
- 537 Res. 67, 216-226.
- 538 Mohamad, N.R., Marzuki, N.H.C., Buang, N.A., Huyop, K. Mahab, R.A., 2015. An
- 539 overview of technologies for immobilization denzymes and surface analysis
- 540 techniques for immobilized enzymes. Biotech ol. Biotec. Eq. 29, 205-220.
- 541 Murugesan, K., Yang, I., Kim, Y., Jun, J Chang, Y., 2009. Enhanced transformation
- 542 of malachite green by laccase of Ganoderma lucidum in the presence of natural

543 phenolic compounds. ppl Microbiol. Biot. 82, 341-350.

544 Prasad, M., Palanivelu, P., 2015. Immobilization of a thermostable, fungal recombinant

- 545 chitinase on biocompatible chitosan beads and the properties of the immobilized
- 546 enzyme. Biotechnol. Appl. Bioc. 62, 523-529.
- 547 Ren, X., Zeng, G., Tang, L., Wang, J., Wan, J., Liu, Y., Yu, J., Yi, H., Ye, S., Deng, R.,
- 548 2018. Sorption, transport and biodegradation-An insight into bioavailability of

- 549 persistent organic pollutants in soil. Sci. Total Environ. 610-611, 1154-1163.
- 550 Sheldon, R.A., van Pelt, S., 2013. Enzyme immobilisation in biocatalysis: why, what
- 551 and how. Chem. Soc. Rev. 42, 6223-6235.
- 552 Shu, Z., Li, T., Zhou, J., Chen, Y., Sheng, Z., Wang, Y., Yuan, X., 2016. Mesoporous
- 553 silica derived from kaolin: Specific surface area enlargement via a new
- zeolite-involved template-free strategy. Appl. Clay Sci. 123, 7
- 555 Shu, Z., Li, T., Zhou, J., Chen, Y., Yu, D., Wang, Y., 2014 Tanylate-free preparation
- 556 of mesoporous silica and alumina from natural molinite and their application in
- 557 methylene blue adsorption. Appl. Clay Sci. 17, 3)-4
- 558 Sinegani, A.S., Emtiazi, G., Shariata adar H., 2005. Sorption and immobilization of
- cellulase on silicate clay mineral. Colloid Interf. Sci. 290, 39-44.
- 560 Sinha, A., Osborne W.L. 016. Biodegradation of reactive green dye (RGD) by
- 561 indigenous fungal strain VITAF-1. Int. Biodeter. Biodegr. 114, 176-183.
- 562 Skoronski, E., Souza, D.H., Ely, C., Broilo, F., Fernandes, M., Junior, A.F., Ghislandi,
- 563 M.G., 2017. Immobilization of laccase from Aspergillus oryzae on graphene
- nanosheets. Int. J. Biol. Macromol. 99, 121-127.
- 565 Sun, H.F., Yang, H., Huang, W.G., Zhang, S.J., 2015. Immobilization of laccase in a

- 566 sponge-like hydrogel for enhanced durability in enzymatic degradation of dye
- 567 pollutants. J. Colloid Interf. Sci. 450, 353-360.
- 568 Tan, D., Yuan, P., Dong, F., He, H., Sun, S., Liu, Z., 2018. Selective loading of
- 569 5-fluorouracil in the interlayer space of methoxy-modified kaolinite for controlled
- 570 release. Appl. Clay Sci. 159, 102-106.
- 571 Tan, X., Liu, Y., Zeng, G., Wang, X., Hu, X., Gu, Y., Yang, 7, 3015. Application of
- 572 biochar for the removal of pollutants from aqueous solutions Chemosphere. 125,
- 573 70-85.
- 574 Tang, W., Zeng, G., Gong, J., Liang, J., Xu, P. Zhang, C., Huang, B., 2014. Impact of
- 575 humic/fulvic acid on the removal of heavy metals from aqueous solutions using
- 576 nanomaterials: A review. Sci. Total Environ. 468, 1014-1027.
- 577 Wan, J., Zeng, G. Huanz, D, Hu, L., Xu, P., Huang, C., Deng, R., Xue, W., Lai, C.,
- 578 Zhou, C., Zheng, K., Ren, X., Gong, X., 2018. Rhamnolipid stabilized
- 579 nano-chlorapatite: Synthesis and enhancement effect on Pb-and Cd-immobilization in
- 580 polluted sediment. J. Hazard. Mater. 343, 332-339.
- 581 Wang, Q., Peng, L., Li, G., Zhang, P., Li, D., Huang, F., Wei, Q., 2013. Activity of
- 582 Laccase Immobilized on TiO<sub>2</sub>-Montmorillonite Complexes. Int. J. Mol. Sci. 14,

583 12520-12532.

- 584 Wang, S., Ning, Y., Wang, S., Zhang, J., Zhang, G., Chen, Q., 2017. Purification,
- 585 characterization, and cloning of an extracellular laccase with potent dye decolorizing
- ability from white rot fungus Cerrena unicolor GSM-01. Int. J. Biol. Macromol. 95,
- 587 <u>920-927</u>.
- 588 Wen, X., Du, C., Zeng, G., Huang, D., Zhang, J., Yin, L., Tan, Fluang, L., Chen, H.,
- 589 Yu, G., Hu, X., Lai, C., Xu, P., Wan, J., 2018. A north prepared by
- 590 immobilized Bacillus licheniformis for lead removation wastewater. Chemosphere.
- 591 200, 173-179.
- 592 Wu, H., Lai, C., Zeng, G., Liang, J., Chen J., Xu, J., Dai, J., Li, X., Liu, J., Chen, M.,
- Lu, L., Hu, L., Wan, J., 2017. Interactions of composting and biochar and their
- 594 implications for soil medment and pollution remediation: a review. Crit. Rev.
- 595 Biotechnol. 37, 754 764.
- 596 Xu, P., Zeng, G.M., Huang, D.L., Feng, C.L., Hu, S., Zhao, M.H., Lai, C., Wei, Z.,
- 597 Huang, C., Xie, G.X., Liu, Z.F., 2012. Use of iron oxide nanomaterials in wastewater
- treatment: A review. Sci. Total Environ. 424, 1-10.
- 599 Xu, P., Zeng, G.M., Huang, D.L., Lai, C., Zhao, M.H., Wei, Z., Li, N.J., Huang, C., Xie,

- 600 G.X., 2012. Adsorption of Pb(II) by iron oxide nanoparticles immobilized
- 601 Phanerochaete chrysosporium: Equilibrium, kinetic, thermodynamic and mechanisms
- 602 analysis. Chem. Eng. J. 203, 423-431.
- 603 Xue, W., Huang, D., Zeng, G., Wan, J., Zhang, C., Xu, R., Cheng, M., Deng, R., 2018.
- Nanoscale zero-valent iron coated with rhamnolipid as an effective stabilizer for
- 605 immobilization of Cd and Pb in river sediments. J. Hazard. Math. 341, 381-389.
- 606 Yang, J., Lin, Y.H., Yang, X.D., Ng, T.B., Ye, X.Y., LA, J. 2017. Degradation of
- 607 tetracycline by immobilized laccase and the proceed transformation pathway. J.
- 608 Hazard. Mater. 322, 525-531.
- 609 Zeng, G., Wan, J., Huang, D., Hu, L., Huang, C., Cheng, M., Xue, W., Gong, X., Wang,
- 610 R., Jiang, D., 2017. Precipitation, casorption and rhizosphere effect: The mechanisms
- 611 for Phosphate-induced Pb/mmobilization in soils-A review. J. Hazard. Mater. 339,612 354-367.
- 613 Zhang, C., Lai, C., Zeng, G., Huang, D., Yang, C., Wang, Y., Zhou, Y., Cheng, M.,
- 614 2016. Efficacy of carbonaceous nanocomposites for sorbing ionizable antibiotic
- sulfamethazine from aqueous solution. Water Res. 95, 103-112.
- 616 Zhang, Y., Zeng, G.M., Tang, L., Chen, J., Zhu, Y., He, X.X., He, Y., 2015.

- 617 Electrochemical Sensor Based on Electrodeposited Graphene-Au Modified Electrode 618 and NanoAu Carrier Amplified Signal Strategy for Attomolar Mercury Detection. 619 Anal. Chem. 87, 989-996. 620 Zheng, F., Cui, B.K., Wu, X.J., Meng, G., Liu, H.X., Si, J., 2016. Immobilization of 621 laccase onto chitosan beads to enhance its capability to degrade synthetic dyes. Int. 622 Biodeter. Biodegr. 110, 69-78. Zhou, C., Lai, C., Huang, D., Zeng, G., Zhang, C., Cheng 623 Wan, J., Xiong, W., Wen, M., Wen, X., Qin, L., 2018. porous carbon nitride by 624 625 supramolecular preassembly of photocatalytic removal of monom for 626 sulfamethazine under visible light Applied Catalysis B: Environmental. 220, lrive 627 202-210. Denizli, A., 2008. Reversible immobilization of catalase 628 Ztrk, N., Tabak, by using a novel bentonite-cysteine (Bent-Cys) microcomposite affinity sorbents. 629 Colloids and Surfaces A: Physicochemical and Engineering Aspects. 322, 148-154. 630
- 631