

1 Synergistic effect of artificial enzyme and 2D nano-structured Bi₂WO₆ for eco-friendly
2 and efficient biomimetic photocatalysis

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11 **Abstract**

12 Highly-efficient and eco-friendly materials and technologies are urgently needed to
13 meet the requirements of nowadays green development. Photocatalysis with using solar
14 energy and enzymatic catalysis with eco-friendly nature are effective alternatives to
15 address the problem. Notably, beneficial use of the synergistic effect of artificial
16 enzyme and advanced photocatalyst has attracted wide attention. This work presents a
17 biomimetic photocatalytic material, two-dimensional (2D) biomimetic hemin-bismuth
18 tungstate (HBWO). Stable HBWO composites formed by immobilization of
19 monomeric hemin on 2D bismuth tungstate layer, exhibit high photocatalytic
20 performance, better than that of pure 2D bismuth tungstate and unsupported hemin.
21 HBWO shows layered structure with the interlayer spacing at ~0.35 nm. In the
22 photocatalytic process, hemin can not only act as an electron shuttle, also play an
23 important role in oxygen transfer. Additionally, the synthesized HBWO composites
24 exhibit nice binding affinities and high photocatalytic activity in tetracycline
25 degradation. It is anticipated that beneficial use of synergistic effect of artificial enzyme
26 and photocatalyst via HBWO composites can be a promising eco-friendly and efficient
27 solution for addressing the environmental crisis.

28

29 **Key words:** Artificial enzyme; Hemin; Two dimension; Bi₂WO₆; Photocatalysis.

30

31

32 1. Introduction

33 Enzymes are well-known to show high catalytic activity, and have been used in
34 addressing energy and environmental crisis for decades [1-10]. However, the practical
35 application of enzymes is limited by the low operational stability, harsh environmental
36 conditions and expensive preparation process [11-16]. To satisfy the needs of the
37 scientific development, the design and promotion of artificial enzymes is developed
38 rapidly [17-21]. Hemin, an attractive biomimetic material, is a promising material for
39 preparation of artificial enzymes, like graphene-hemin, G-quadruplex/hemin and
40 hemin-AO-PAN [22-24]. Up to date, hemin has been well studied as catalysts because
41 of their unique physicochemical behavior, especially high conductivity [25-28]. The
42 catalytic performance of hemin is largely influenced by the electron transfer between
43 the central iron and substances [29]. Hemin can play an important role in electron
44 transfer in the catalytic process. In addition, hemin has been proved to be an oxygen
45 transfer reagent [30, 31]. Nevertheless, monomeric hemin in solution is easy to
46 aggregate to form inactive dimers, which has a negative effect in the catalytic
47 performance. Efficient methods to maintain the activity of monomeric hemin are
48 urgently needed.

49 In recent years, many methods have been developed, especially the
50 immobilization process [32-35]. For example, Huang et al. [36] used graphene as the
51 support to immobilize hemin to be a highly biomimetic oxidation catalyst, showing
52 high catalytic activity for the pyrogallol oxidation reaction with H₂O₂. Pyrogallol was
53 oxidized to purpurogallin with catalytic efficiency at $2.0 \times 10^5 \text{ M}^{-1} \text{ min}^{-1}$. Yao et al. [29]

54 introduced multi-walled carbon nanotubes to improve the catalytic performance of
55 hemin on methylene blue degradation with the assistance of H₂O₂. The elimination
56 kinetics reached 0.04195 min⁻¹. Lu et al. [37] combined g-C₃N₄ with hemin via axial
57 coordination to degrade 4-chlorophenol, which not only maintained the stability of
58 hemin but also improved the photocatalytic performance of g-C₃N₄, but it is worth
59 noting that 4-chlorophenol degradation efficiency was reduced 72% without H₂O₂
60 compared with that in the presence of H₂O₂. H₂O₂ is still needed in these efficient
61 catalytic processes, which is uneconomical for practical application. How to solve the
62 limitation of the advanced biomimetic hemin-contained catalyst is a challenge.

63 Since graphene (Nobel Prize in 2010) became the research hotspot, studies on
64 two-dimensional (2D) nanomaterials have escalated sharply due to their excellent
65 photoelectric properties [38-44]. 2D Bismuth tungstate (BWO), constructed by [WO₄]²⁻
66 - [Bi₂O₂]²⁺ - [WO₄]²⁻ layers, is one of the most promising photocatalytic materials [45,
67 46]. The layered structure benefits the photogenerated charge-carriers separation
68 [47-53]. Notably, compared with one-dimensional and three-dimensional structure, 2D
69 BWO is believed to perform better photocatalytic performance: (i) charge migration
70 distance can be reduced for the bulk to the surface [54]; (ii) the uncoordinated surface
71 atoms in 2D structure contribute to the higher harvest of solar light [55, 56]; (iii) large
72 open and accessible surface area benefits the substrate diffusion and binding
73 interactions [57, 58], also allows higher light absorption and more oxygen vacancies
74 generation [59-61]; (iv) 2D structure is an excellent platform for the design of novel
75 multicomponent materials to achieve higher efficient catalytic performance [62, 63].

76 Therefore, in this work, we used 2D BWO as the support for hemin to maintain
77 their activity, meanwhile the introduction of hemin showed the potential to promote the
78 photocatalytic activity of 2D BWO by improving the electronic property. Monomeric
79 hemin was immobilized on the surface of 2D BWO to form hemin-bismuth tungstate
80 (HBWO) composites via a facile hydrothermal method. The structure and photoelectric
81 property were explored in detail. The photocatalytic performance of HBWO
82 composites were investigated by the degradation of tetracycline (TC), which commonly
83 exist in environment and are harmful for living beings. It is anticipated that the 2D
84 nano-structured HBWO composite can be a promising hemin-based photocatalytic
85 material for the degradation of environmental pollutants via an eco-friendly way.

86 **2. Experimental**

87 2.1. Synthesis of catalysts

88 All the reagents were analytical grade. HBWO were prepared through one-step
89 hydrothermal process. Typically, $\text{Bi}(\text{NO}_3)_3 \cdot 5\text{H}_2\text{O}$ was dispersed in 6.8 wt% nitrate
90 solution to get solution A with concentration at 0.067 M, and $\text{Na}_2\text{WO}_4 \cdot 2\text{H}_2\text{O}$ was
91 dispersed in deionized water to get solution B (0.033 M). Then solution B was added
92 dropwise into the solution A under stirring, mixed with 0.05 g of cetyltrimethyl
93 ammonium bromide and a certain amount of hemin methanol solution. After
94 magnetically stirring for 6 h and adjusting pH to 7, the mixture was transferred to
95 Teflon-lined autoclave and sealed into a steel tank, then heated to 180 °C and
96 maintained for 16 h. 2D HBWO was obtained. The different mass ratio of hemin/BWO
97 at 0.5 wt%, 1.0 wt%, 3.0 wt%, 5.0 wt%, 7.0 wt% were prepared and signed as

98 0.5HBWO, 1HBWO, 3HBWO, 5HBWO, 7HBWO, respectively. 2D BWO was
99 synthesized via the parallel preparation process without hemin.

100 2.2. Characterization

101 Crystallographic information was recorded by powder X-ray diffraction
102 (XRD-6100, Cu $K\alpha$ radiation, $\lambda=0.15418$ nm). The morphology and microstructure
103 were investigated by high resolution transmission electron microscopy (HRTEM,
104 Tecnai G2 F20 S-TWIN) and atomic force microscopic (AFM, Bioscope system). The
105 surface elemental compositions were identified by X-ray photoelectron spectroscopy
106 (XPS, ESCALAB 250Xi spectrometer).

107 2.3. Photocatalytic experiments

108 Tetracycline (TC) was chosen as the target pollutant. The photocatalytic activities
109 were investigated by the photodegradation of TC under simulated solar light irradiation
110 (Xe lamp, CELHXF300). 0.2 g/L catalyst was dispersed into 10 mg/L TC solution in
111 the photodegradation experiment. The Xiangjiang River water, Taozi Lake water and
112 tap water (Changsha, Hunan, China) were filtrated by a 0.22 μm filter membrane.
113 Before irradiation, dark reaction was performed to reach the adsorption/desorption
114 equilibrium. Reaction solution was collected at given time interval and then filtered
115 through 0.45 μm membrane filters. TC concentration was analyzed by a Shimadzu
116 UV-vis spectrophotometer with the absorbance at the characteristic band of 357 nm.
117 The photodegradation intermediates of TC were obtained by a high-performance liquid
118 chromatography-mass spectrometry (HPLC-MS, 1290/6460 Triple Quad). Each
119 experiment for photocatalytic activities was performed in triplicate.

120 2.4. Optical property and photo-electrochemical measurements

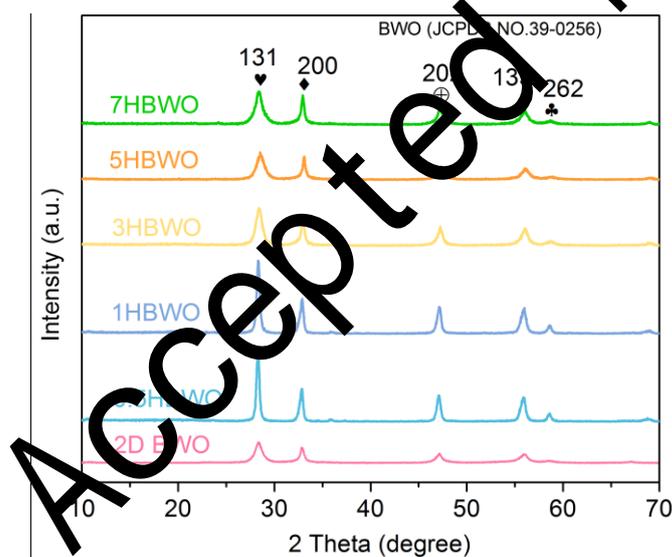
121 Optical properties were investigated by UV-vis diffuse reflectance
122 spectrophotometer (DRS, Hitachi U4100 UV), photoluminescence measurements (PL,
123 FluroMax-4), and transient photocurrent response, Mott-Schottky and electrochemical
124 impedance spectroscopy (IT, M-S and EIS, CHI760E workstation). The electron spin
125 resonance (ESR) signals of radicals spin-trapped by 5,5-dimethyl-1-pyrroline N-oxide
126 (DMPO) in water were performed on a Bruker ER200-SRC spectrometer. And the
127 effect of h^+ , $\bullet O_2^-$ and $\bullet OH$ in TC degradation was detected by active species trapping
128 tests. Ethylenediaminetetraacetic acid disodium salt (EDTA-2Na),
129 tetramethylpiperidine (TEMPOL) and isopropyl alcohol (IPA) were used as the
130 scavengers for h^+ , $\bullet O_2^-$ and $\bullet OH$, respectively.

131 3. Results and discussion

132 3.1. Characterizations

133 XRD analysis was used to record the crystal phase. As presented in Fig. 1, the
134 distinct diffraction peaks of all the samples located at 28.3° , 32.8° , 47.1° , 56.0° , and
135 58.5° , corresponding to the (131), (200), (202), (133), (262) crystal planes of
136 orthorhombic BWO (JCPDS NO. 39-0526), respectively [55]. No other crystalline
137 phase was detected, suggesting that the introduction of hemin did not affect the crystal
138 phase of 2D BWO. The peak intensity of HBWO composites was higher than that of
139 2D BWO, showing that the introduction of moderate hemin improved the
140 crystallization. The increased peak intensity of HBWO might be because the crystal
141 nucleation centers preferred on hemin owing to the aromatic system, and then HBWO

142 with larger particle size than BWO was formed (Fig. 3) [64-66]. However, the intensity
 143 of the peaks decreased with the hemin weight ratio increased, showing that excess
 144 hemin restrained the crystallization course of BWO. This because it is easier for
 145 excess hemin to form hemin aggregates, which might affect the role of aromatic
 146 system in the crystallization of HBWO [67]. Based on these results, it was preliminary
 147 concluded that optimum mass ratio of hemin/BWO was 1.0 wt%. Additionally,
 148 according to Bragg equation ($2d\sin\theta = n\lambda$, d is the interplanar spacing, θ is the Bragg
 149 angle, n is the order of the reflection, and λ is equal to 0.15418 nm), 0.32 nm of the
 150 interplanar spacing of 1HBWO can be obtained.

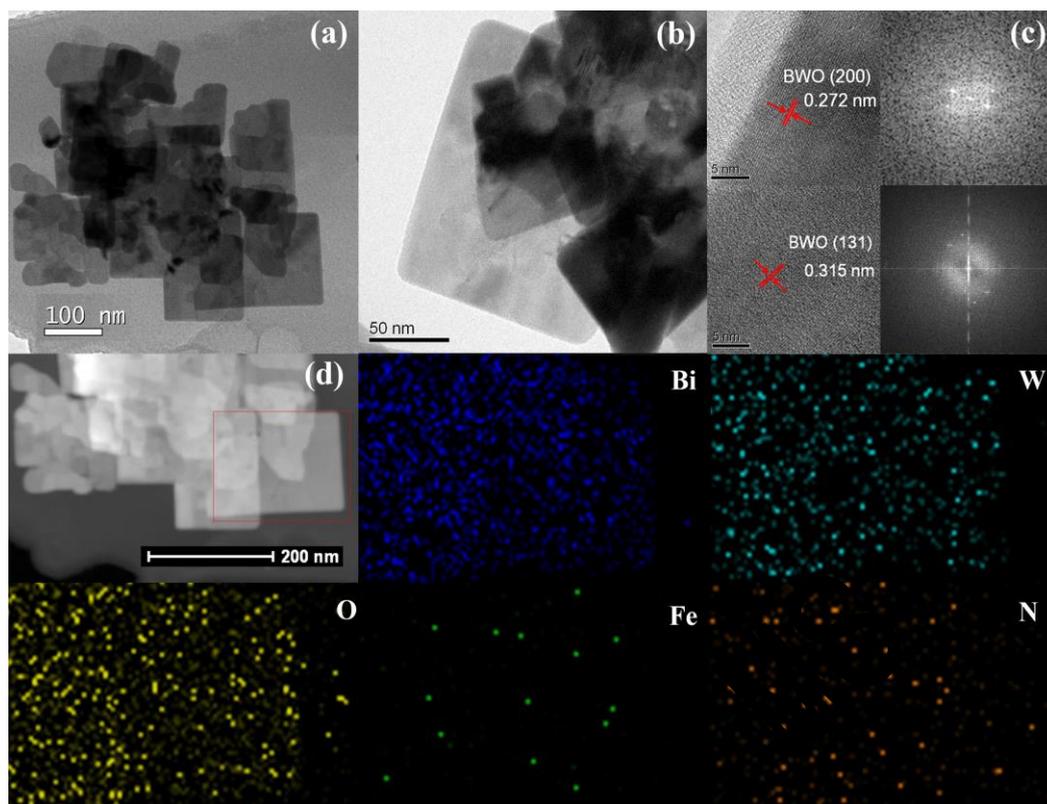


151
 152 Fig. 1. XRD analysis of 2D BWO and HBWO composites.

153 Morphology and microstructure of 1HBWO were observed by HRTEM. As shown
 154 in Fig. 2, Fig 2a and b exhibited the prepared 1HBWO of a regular sheet shaped
 155 structure with size of $\sim 150 \times 250\text{ nm}$. The d spacing value of the lattice fringe, 0.272
 156 nm and 0.315 nm , were found in Fig. 2c, corresponding to the (200) and (131) planes,
 157 respectively (Table S1). The composition of 1HBWO was further explored via

158 high-angle annular dark-field scanning transmission electron microscopy
159 (HAADF-STEM) and energy dispersive X-ray spectroscopy (EDS) elemental scanning.
160 The results were presented in Fig. 2d, showing the distribution of Bi, W, O, Fe, and N
161 elements in 1HBWO. And the homogeneous distribution of Fe and N originating from
162 hemin can be found obviously.

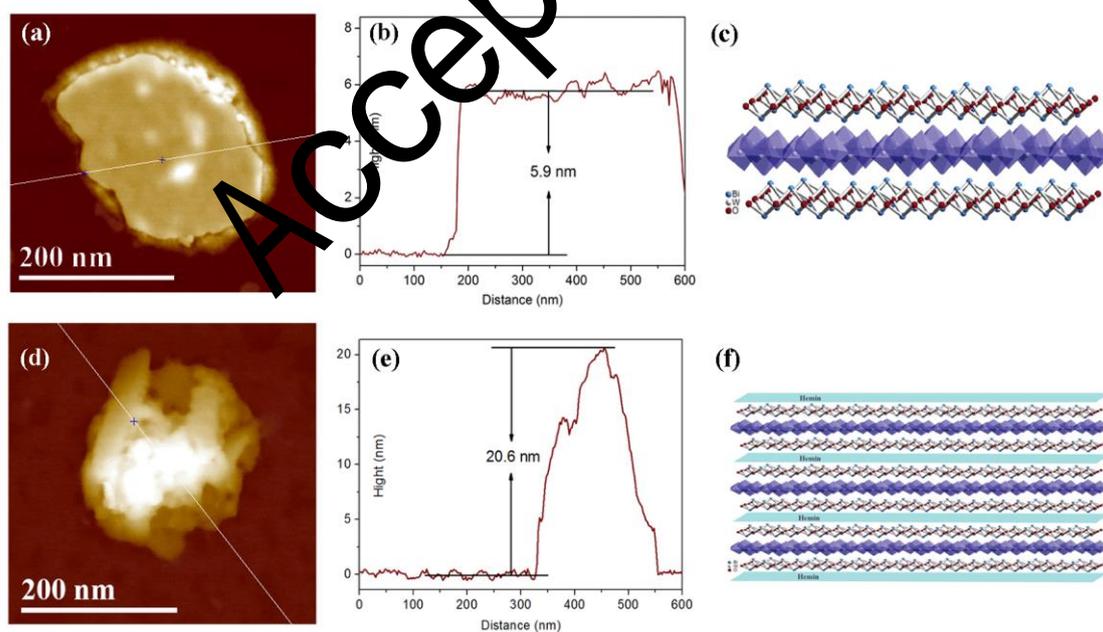
163 Detailed size and thickness were determined by AFM, shown by the AFM images
164 and the corresponding height histograms. As-prepared pure 2D BWO showed a size of
165 $\sim 180 \times 200$ nm and a thickness of 5.9 nm (Fig. 3a, b, c), while 1HBWO showed a size
166 of $\sim 200 \times 200$ nm and a thickness of 20.6 nm (Fig. 3d, e). Presumably, monomeric
167 hemin with the thickness of 0.2 nm [36] was uniformly assembled on 2D BWO to form
168 1HBWO nanomaterials, which might show a sandwich structure built by
169 hemin-BWO-hemin layers (Fig. 3f). We speculated that 1HBWO contained 3~4 layers
170 with a suitable interlayer spacing. A numerical equation can be built: $A + m(B+A) +$
171 $2m \times D = C$ (A is the thickness of monomeric hemin, B is the thickness of BWO, C is
172 the thickness of 1HBWO, D is the interlayer spacing, m is the number of layers). D
173 (the interlayer spacing) is calculated to be 0.35 nm. Overall, HRTEM and AFM
174 analysis affirmed the successful assembly of 2D nano-structured 1HBWO.



175

176 Fig. 2. (a, b, c) TEM and HR-TEM images of 1HBWO; (d) HAADF-STEM image and the corresponding

177 EDS elemental mapping images of 1HBWO composite.

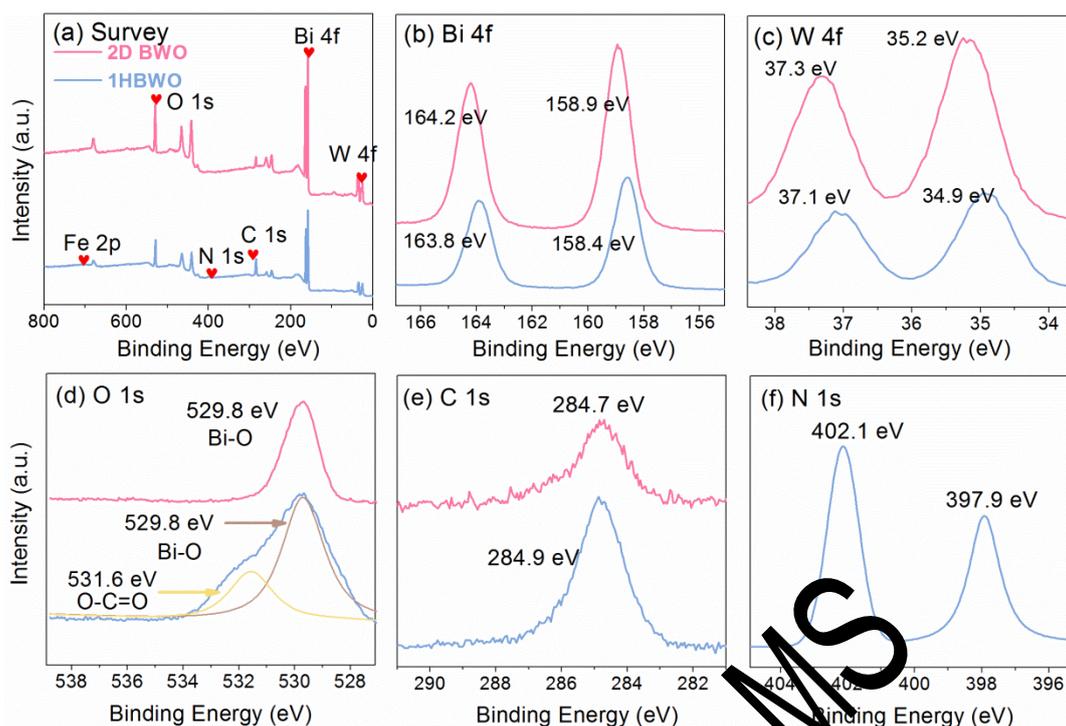


178

179 Fig. 3 (a) AFM image, (b) cross-section profile and (c) proposed structural diagram of 2D BWO; (d)

180 AFM image, (e) cross-section profile and (f) proposed structural diagram of 1HBWO composite.

181 The surface elemental composition and oxidation state of 1HBWO were
182 determined by XPS. Fig. 4a represented a survey spectrum, indicating the main peaks
183 of Bi, W, C and O element. The 1HBWO sample showed Bi 4f and W 4f signals with a
184 Bi/W ratio of 2.02, which is nearly equal to stoichiometric ratio of 2D BWO.
185 High-resolution spectra of Bi 4f (Fig. 4b) at 158.4 eV and 163.8 eV (5.4 eV of doublet
186 separation energy) corresponded to the binding energy of Bi 4f_{7/2} and Bi 4f_{5/2},
187 respectively, indicating the existence of Bi³⁺. Correspondingly, high resolution spectra
188 of W 4f (Fig. 4c) at 34.9 eV and 37.1 eV (2.2 eV of /doublet separation energy) belong
189 to the binding energy of W 4f_{7/2} and W 4f_{5/2} with a ratio of 4:3 in peak area, indicating
190 the same valence state of W in +6 oxidation state. Compared with the peak of BWO,
191 the Bi 4f peak of 1HBWO displayed a shift ~0.4 eV and the W 4f peak showed a shift
192 ~0.3 eV to lower binding energies because of the interaction between hemin and BWO.
193 The O 1s peaks at 529.8 eV and 531.6 eV (Fig. 4d) belong to the binding energy of Bi-O
194 and O-C=O. The C 1s peak at 284.9 eV (Fig. 4e) is attributed to C-C/-CH_x [68]. In the N
195 1s spectra (Fig. 4f), binding energy peak of C=N-C is located at 397.9 eV [69], and the
196 peak at 402.1 eV is originating from pyridine-N [70]. XPS analysis indicates the
197 presence of hemin and BWO in 1HBWO.



198

199 Fig. 4. XPS analysis of 2D BWO and 1HBWO composite: (a) survey spectra, (b) Bi, (c) W 4f; (d) O 1s;

200 (e) C 1s; (f) N 1s.

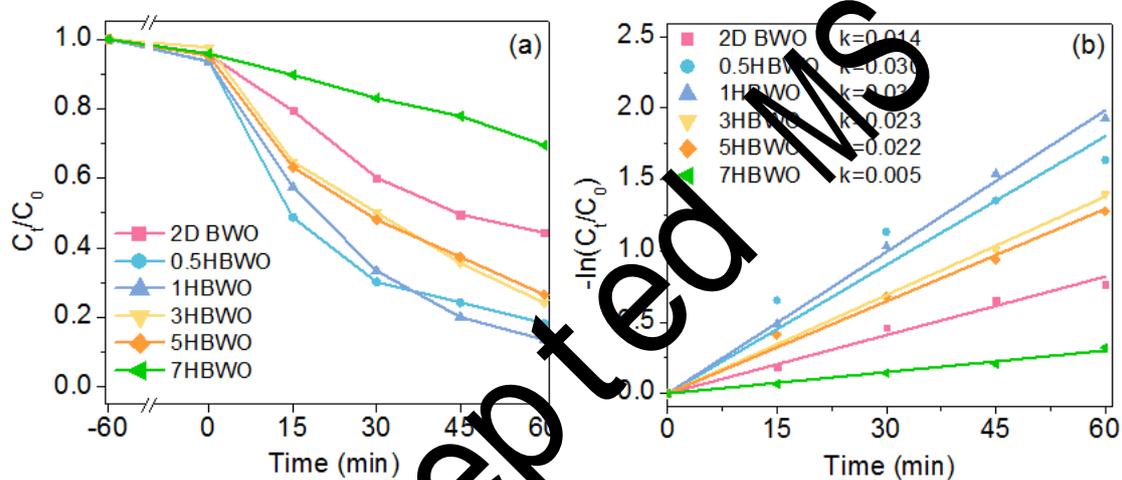
201 3.2. Photoactivity test

202 TC, a representative antibiotic is widely used by many industries during the last
 203 decades. In this work, the photoactivity of HBWO composites was determined by the
 204 photocatalytic degradation of TC under simulated solar light irradiation, exploring the
 205 effect of different mass ratio of hemin/BWO, the effect of water matrix and the effect of
 206 inorganic salt ions. According to the dark adsorption experiment of HBWO, the
 207 absorption of TC reached absorption equilibrium within 60 min. As shown in Fig. S1,
 208 after 60 min, 4.5%, 6.3%, 6.2%, 2.4%, 4.8% and 4.2% of TC were absorbed by 2D
 209 BWO, 0.5HBWO, 1HBWO, 3HBWO, 5HBWO, and 7HBWO, respectively. The
 210 absorption experiment demonstrated that the introduction of hemin had a small impact
 211 on the absorption of TC.

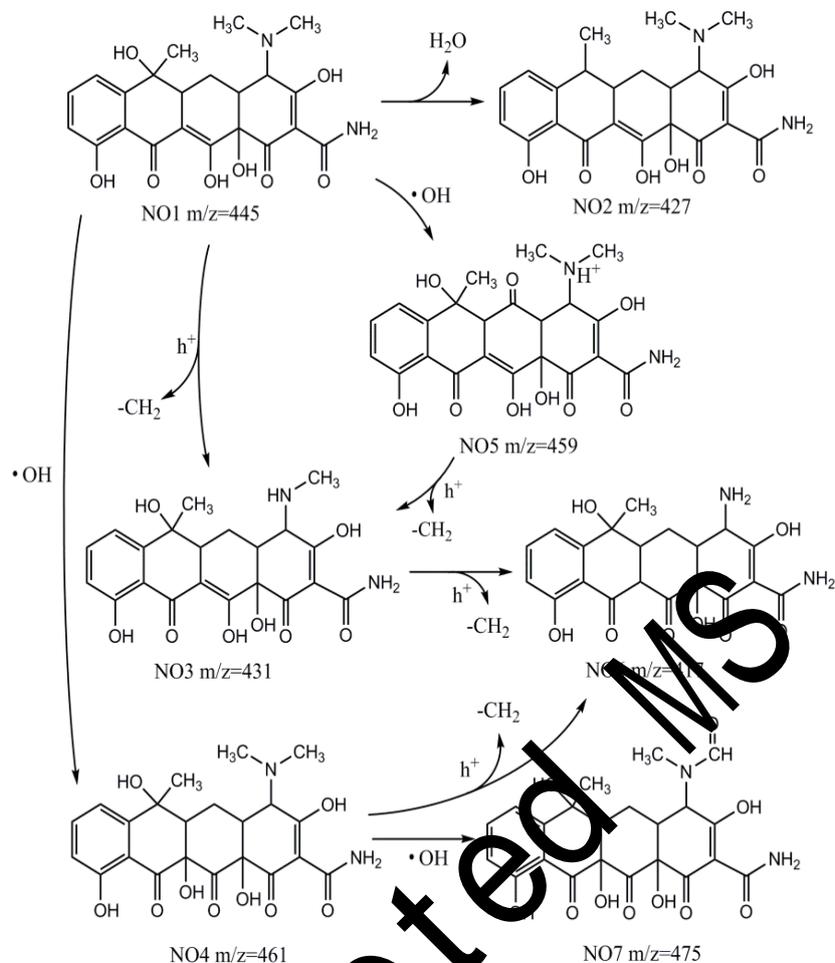
212 Fig. 5a and b showed the photocatalytic degradation efficiency of TC over HBWO
213 with different mass ratio of hemin/BWO. As shown in Fig. 5a, 55.6% TC can be
214 photodegraded over 2D BWO after 60 min. When 0.5%-5.0% hemin was introduced,
215 the photocatalytic performance of TC degradation was improved. The 1HBWO
216 exhibited the highest photocatalytic efficiency, which reached 86.4%. The
217 photocatalytic degradation efficiency of HBWO composites decreased with increasing
218 hemin weight ratio (when ratio > 1.0 wt%). 7HBWO showed the worst photocatalytic
219 performance on TC degradation. This might because too much hemin immobilized on
220 the surface of 2D BWO might shadow the reactive sites. Overall, the introduction of
221 hemin with the content range at 0.5%-5.0% can enhance the photocatalytic
222 performance of HBWO. The photocatalytic degradation kinetics of TC was also
223 explored, which was shown in Fig. 5b. Obviously, 1HBWO showed the highest rate
224 constant of 0.033 min^{-1} . 1HBWO was the optimum photocatalyst, which was used in
225 the next photocatalytic tests. And the catalytic performance of HBWO was compared
226 with several reported bimimetic catalysts, listed in Table S2.

227 Additionally, the photodegradation intermediates of TC were detected by
228 HPLC-MS. HPLC-MS spectra of the TC intermediates eluted in the photodegradation
229 process were presented in Fig. S2 and the intermediates were listed in Table S3. It can
230 be seen from Fig. S2 that the intensity of the peak with m/z of 445 (TC) was decreasing
231 with the reaction time. Dehydration and hydroxylation process by $\bullet\text{OH}$ attack and
232 N-demethylation process (low bond energy of N-C) via h^+ attack played key roles in the
233 TC photodegradation [71, 72]. The possible degradation processes were presented in

234 Fig. 6. Intermediate NO₂ was generated via the dehydration process. •OH can directly
 235 attack the TC molecules in the hydroxylation process to produce intermediate NO₄ and
 236 NO₅, and then NO₄ can be degraded to NO₇. Intermediate NO₃ was formed mainly via
 237 the N-demethylation process of TC and NO₅. Additionally, intermediate NO₆ can be
 238 produced from the N-demethylation of NO₃ and NO₄. Overall, these detected
 239 intermediates confirmed that TC was degraded progressively via the photocatalytic
 240 process over HBWO.



241
 242 Fig. 5. (a) Photocatalytic degradation efficiency and (b) kinetics (k / min^{-1}) of TC over 2D BWO and
 243 HBWO composites with different mass ratio of hemin/BWO.



244

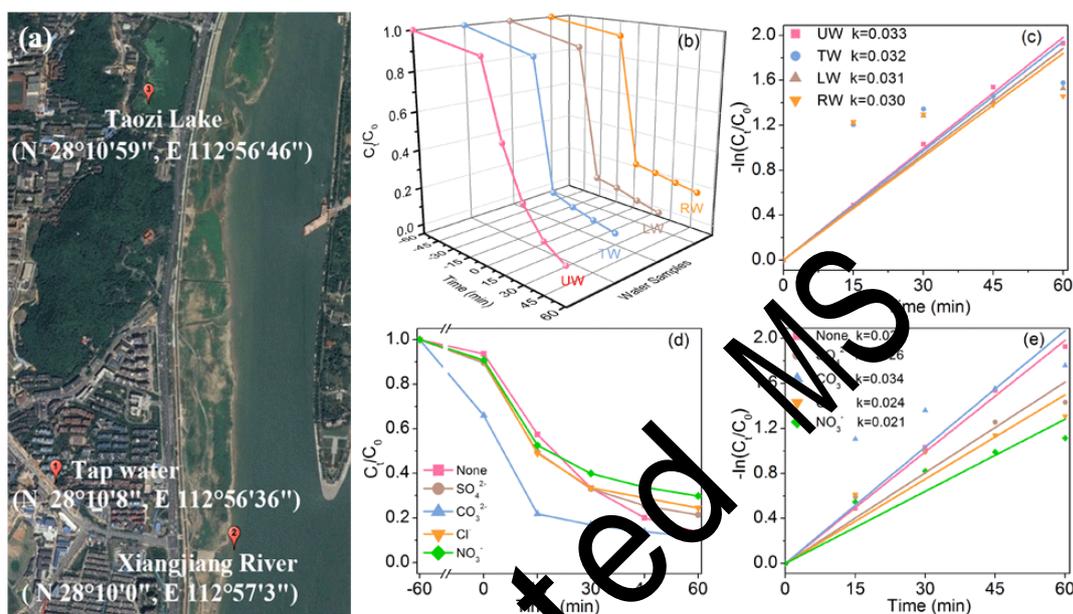
245 Fig. 6. Photodegradation intermediates of TC and the possible degradation processes.

246 In practical application, water matrix is a crucial factor. In this work, ultrapure
 247 water (UW), tap water (TW), Lake water (LW) and River water (RW) were used to
 248 investigate the effect of water matrix on the photocatalytic degradation of TC over
 249 1HBWO. The sampling sites were shown in Fig. 7a. After 60 min irradiation,
 250 photocatalytic efficiency of TC degradation achieved 86.4%, 81.4%, 80.2%, and 78.1%
 251 (Fig. 7b), and the rate constant was 0.033 min⁻¹, 0.032 min⁻¹, 0.031 min⁻¹, and 0.030
 252 min⁻¹ (Fig. 7c) in TC-obtained UW, TW, LW, and RW, respectively. According to the
 253 detected parameters of the water sources (Table 1), the pH value increased in the order
 254 as follows: RW<LW<TW. Based on the experimental results, presumably a slight

255 higher pH range at 6.0-7.5 had a positive effect on the photocatalytic degradation of TC
256 over 2D 1HBWO in natural water. This positive effect might be ascribed to the surface
257 charge variations of TC adsorbed onto the surface of catalysts [73]. Photodegradation
258 of TC over 1HBWO composite maintained high efficiency in real water matrix.

259 Besides, the effect of inorganic salt ions (including CO_3^{2-} , Cl^- , SO_4^{2-} and NO_3^-)
260 was investigated, because the electrolytes might have an effect on the photocatalytic
261 activity. As shown in Fig. 7d and e, Cl^- , SO_4^{2-} and NO_3^- had a little negative effect on
262 the photocatalytic efficiency of TC degradation over 1HBWO composite with the
263 solution pH at about 6.0. The negative effect can be listed in the following decreasing
264 order: $\text{NO}_3^- > \text{Cl}^- > \text{SO}_4^{2-}$. According to the Zeta potentials of the TC solution (Table 2),
265 the Zeta potentials of Cl^- -TC solution, SO_4^{2-} -TC solution and NO_3^- -TC solution were
266 more positive than that of the condition (-29.8 ± 1.5), showing that TC adsorption on
267 the catalysts might be inhibited. The decrease in the photocatalytic efficiency might
268 attribute to the enhanced electrostatic repulsion and the competitive adsorption between
269 the inorganic ions and TC for the limited active sites on the catalysts [74]. The Zeta
270 potential of SO_4^{2-} -TC solution (-23.6 ± 1.8) was closest to that of the condition (-29.8
271 ± 1.5), contributing to the lowest negative effect caused by SO_4^{2-} . NO_3^- showed
272 photoabsorption at the solar light ranging from 290 - 370 nm, leading to the
273 light-filtering effects, which negatively affected the photodegradation efficiency [75].
274 This might be the reason why NO_3^- showed higher negative effect than Cl^- . On the
275 contrary, CO_3^{2-} slightly facilitated the photocatalytic degradation of TC over 1HBWO
276 composite. This might be ascribed to the enhanced adsorption of TC on catalysts owing

277 to the more negative Zeta potential (-31.9 ± 0.8) and higher pH value (9.64) caused by
 278 the generation of OH^- [73]. Overall, the results demonstrated that the photocatalytic
 279 process over 1HBWO composite was an efficient technology for TC removal in
 280 practical application.



281
 282 Fig. 7. (a) The sampling sites of natural water sources; (b) Photocatalytic degradation and (c) kinetics of
 283 TC over 1HBWO composite in ultra-pure water (UW), tap water (TW), Lake water (LW) and River
 284 water (RW) (k/min^{-1}); (d) Effect of coexisting inorganic salt ions on the degradation efficiency and (e) kinetics
 285 of TC over 1HBWO (k/min^{-1}): 0.5 mM of coexisting anions in 10 mg/L TC solution.

286 **Table 1** Parameters of the tap water (TW), Lake water (LW) from Taozi Lake, and River water (RW)
 287 from Xiangjiang River.

Samples	pH	Cl^- (mg L^{-1})	SO_4^{2-} (mg L^{-1})	NO_3^- (mg L^{-1})
TW	7.15	0.058	16.2	1.260
LW	6.91	0.074	14.3	0.872
RW	6.72	0.084	20.5	1.130

288 **Table 2** Zeta potentials and pH of the TC solution in the presence of HBWO with different inorganic salt
289 ions.

Inorganic salt ions	None	CO ₃ ²⁻	Cl ⁻	SO ₄ ²⁻	NO ₃ ⁻
Zeta potential (mV)	-29.8 ± 1.5	-31.9 ± 0.8	-10.2 ± 1.7	-23.6 ± 1.8	-17.8 ± 0.9
pH	5.68	9.64	6.10	5.97	5.95

290 a. Experimental conditions: m/v = 0.2g/L, TC (100ml, 10mg/L), and inorganic salt ions concentration
291 (0.5mM).

292 3.3. Photocatalytic mechanism

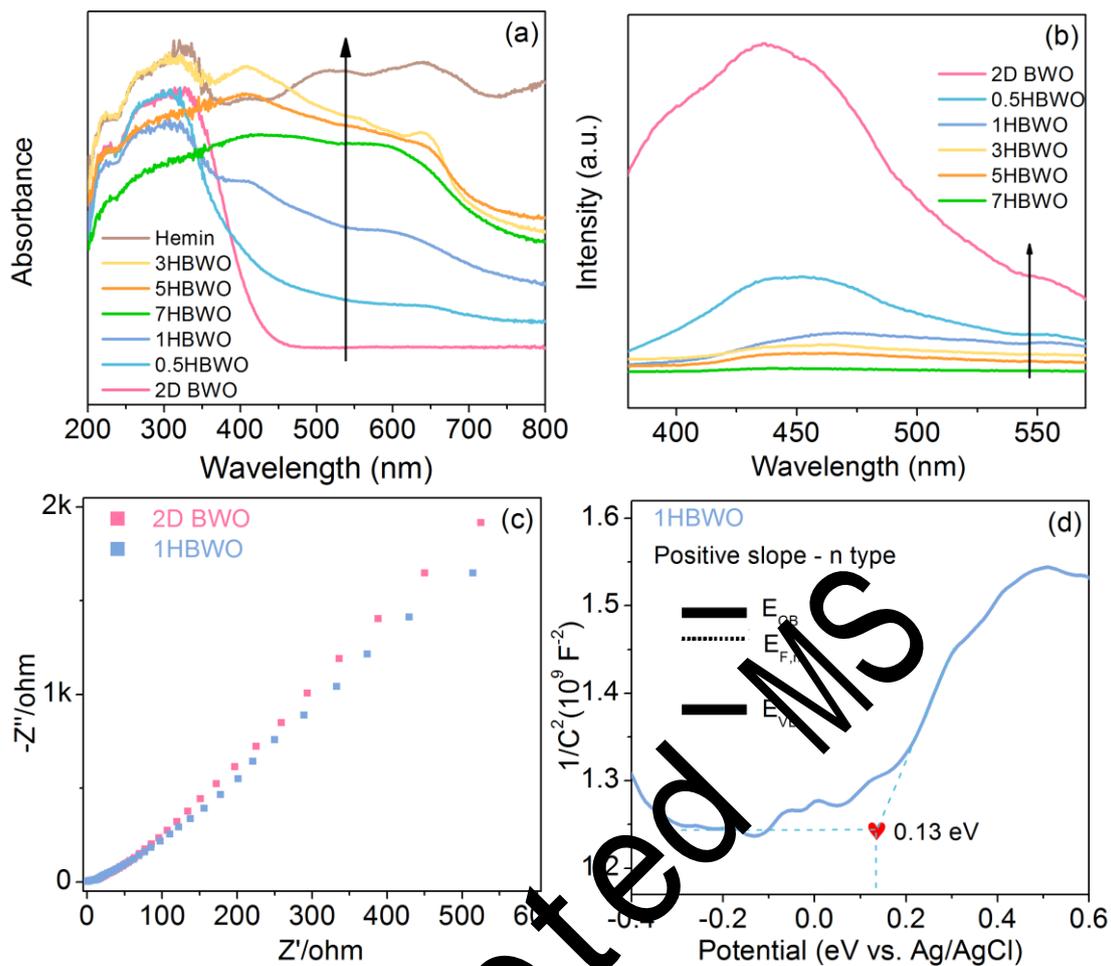
293 3.3.1 Optical property

294 The optical property of pure 2D BWO and HBWO composites were investigated
295 by UV-vis DRS, which were shown in Fig. 7a. It can be found that pure 2D BWO
296 exhibits the absorption of light shorter than 450 nm, and HBWO composites show
297 enhanced optical absorption in the range 450-800 nm. Hemin showed the highest
298 absorption of light owing to the blackbody nature. This phenomenon suggested that
299 HBWO composites might possess higher photoactivity, and hemin played an important
300 role in the enhanced photoabsorption. Fig. 8b showed steady-state PL spectra of
301 as-prepared photocatalysts, which indicated the behavior of photogenerated charge
302 carriers [76]. It can be found that the emission peak centered at around 450 nm, and
303 HBWO composites showed a lower intensity of centered peak, indicating the reduced
304 recombination of photoelectrons and holes. However, introduction of too much hemin,
305 like 7HBWO, might form inactive hemin aggregates and shadow the surface active

306 sites of 2D BWO photocatalysts, and then reduce the generation of charge carriers. The
307 introduction of hemin in moderation can not only enhance the utilization of wider
308 spectrum light irradiation, also boost the separation of photogenerated electrons and
309 holes, thus improving the photocatalytic performance.

310 In order to further prove this assumption, EIS and IT were performed under light
311 irradiation. From the EIS Nyquist plot (Fig. 8c), it can be found that the arc radius of
312 the ITO/1HBWO film was lower than that of the ITO/2D BWO film, exhibiting a
313 higher efficiency of electron transfer at the electrodes. The electron transfer efficiency
314 is affected by the corresponding charge transport behavior [77]. The photogenerated
315 electrons were reduced by the oxygen carried by 1HBWO, accelerating the charge
316 carriers transfer and then showing a higher electron-transfer efficiency. According to
317 the IT result (Fig. S3), 1HBWO showed the increased transient photocurrent than 2D
318 BWO, indicating the improved separation of electrons and holes.

319 Fig. 8d showed M-S plot of 1HBWO, which showed a positive slope, indicating
320 the behavior of *n*-type semiconductor [78]. So the introduction of hemin did not change
321 the *n*-type behavior of 2D BWO. In *n*-type semiconductor, the bottom of the conduction
322 band (CB) is more negative (about 0.2 eV) than Fermi level. Besides, M-S analysis can
323 be used to determine the flat band potential via extrapolating to the intercept of the
324 x-axis. The flat band potential was inferred to be around +0.13 eV vs. Ag/AgCl. And
325 generally, the Fermi level is equal to the flat band potential. Therefore, the E_{CB} of
326 HBWO can be inferred to be around +0.13 eV vs. NHE, which is more negative than
327 $E^0(O_2/H_2O_2)$ (+0.682 V vs. NHE).



328

329 Fig. 8. Optical property analysis: (a) DRS of the prepared catalysts, (b) PL of the prepared catalysts, and

330 (c) EIS of 2D BWO and 1HBWO; (d) M-S analysis of 1HBWO film electrode in 0.5M of Na₂SO₄

331 solution with pH locked at 5.06.

332 3.3.2 Roles of radical species

333 To investigate the presence of radical species in the degradation of TC over

334 HBWO composite, ESR was conducted. As shown in Fig. 9a and b, both •OH and •O₂⁻

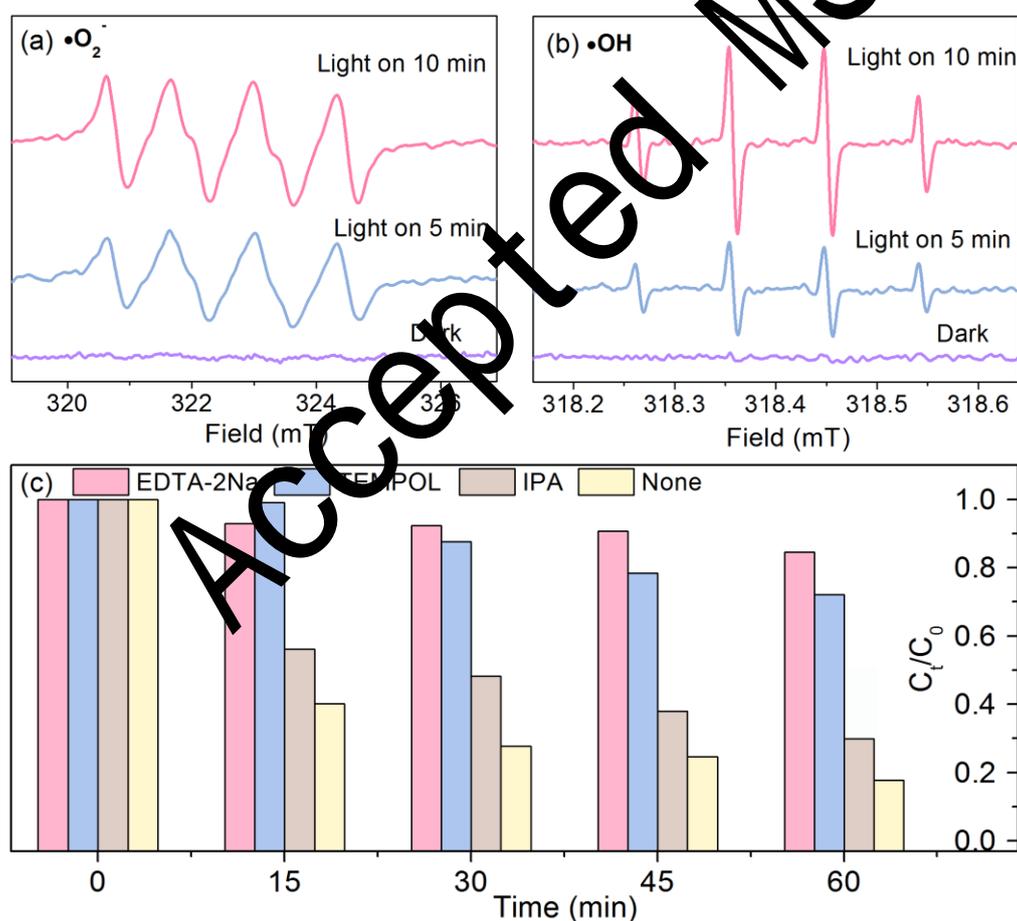
335 radicals signal were observed after irradiation, suggesting the generation of •OH and

336 •O₂⁻ radicals during the photocatalytic process, which can improve the photocatalytic

337 performance of HBWO composite. After that, scavenger tests were performed to study

338 the roles of active radical species in TC photodegradation. Fig. 9c showed that the

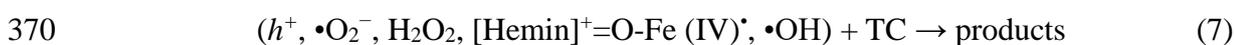
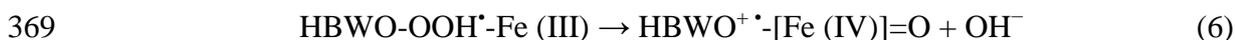
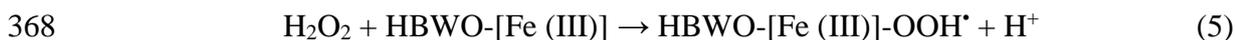
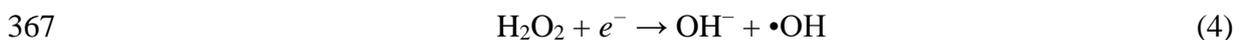
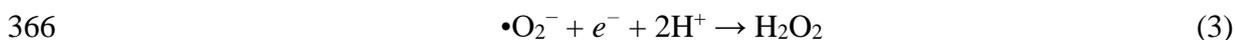
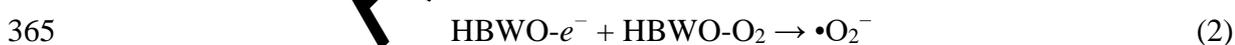
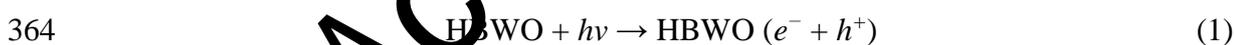
339 photocatalytic efficiency of TC degradation was inhibited greatly when EDTA-2Na and
 340 TEMPOL was added, suggesting that holes and $\bullet\text{O}_2^-$ played the main roles in the
 341 catalytic process. The degradation efficiency decreased slightly when adding IPA,
 342 suggesting that just a small amount of $\bullet\text{OH}$ radicals participated in photocatalytic
 343 degradation of TC. The ESR analysis and scavenger tests indicated that photogenerated
 344 holes and $\bullet\text{O}_2^-$ were the main active species in the photocatalytic degradation of TC.
 345 Additionally, maximum 36 $\mu\text{mol/L}$ of H_2O_2 was generated in the photocatalytic process,
 346 measured via UV-vis spectrophotometer after complexation with molybdate [79].

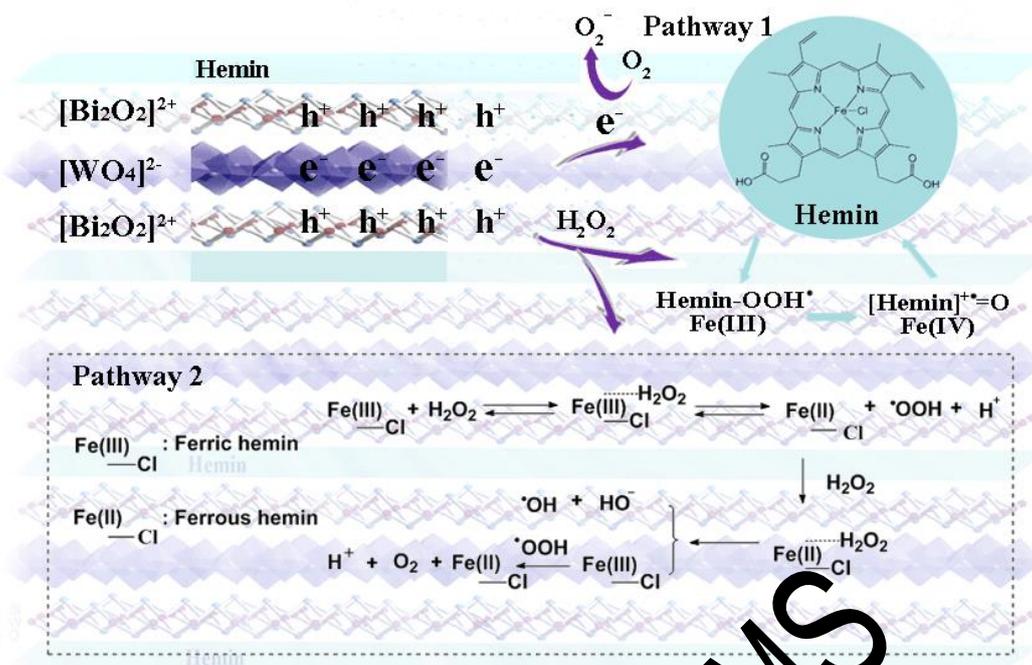


347
 348 Fig. 9. ESR spectra of (a) DMPO•O₂⁻ adduct in methanol dispersion and (b) DMPO•OH adduct in
 349 aqueous dispersion for 1HBWO; (c) scavenger tests: 2 mmol scavengers in 100 mL 10mg/L TC solution.

350 3.3.3 Mechanism of the improved photocatalytic activity

351 Based on the above results, the improved photocatalytic performance of HBWO
352 composites can be explained as follows (Fig. 10): (i) the introduction of moderate
353 hemin in HBWO can enhance the utilization of wider spectrum of light owing to the
354 blackbody nature; (ii) 2D structure benefited the faster transfer of photogenerated
355 charge carriers, and the uncoordinated surface atoms in the 2D structure can improve
356 the harvest of solar light; (iii) hemin played as an electron shuttle to transfer the
357 electrons in CB of HBWO, thus enhanced the separation efficiency of electrons and
358 holes. The holes left in VB can oxidize TC directly; (iv) hemin is an oxygen transfer
359 reagent, and the oxygen molecules carried by hemin can combined with the electrons
360 on CB of HBWO to form $\bullet\text{O}_2^-$, leading to the degradation of TC; (v) H_2O_2 generated in
361 the photocatalytic process can induce the Fenton-like reaction (two possible pathways)
362 with high oxidizing capacity [37, 68]. The photocatalytic degradation process can be
363 described as follows:



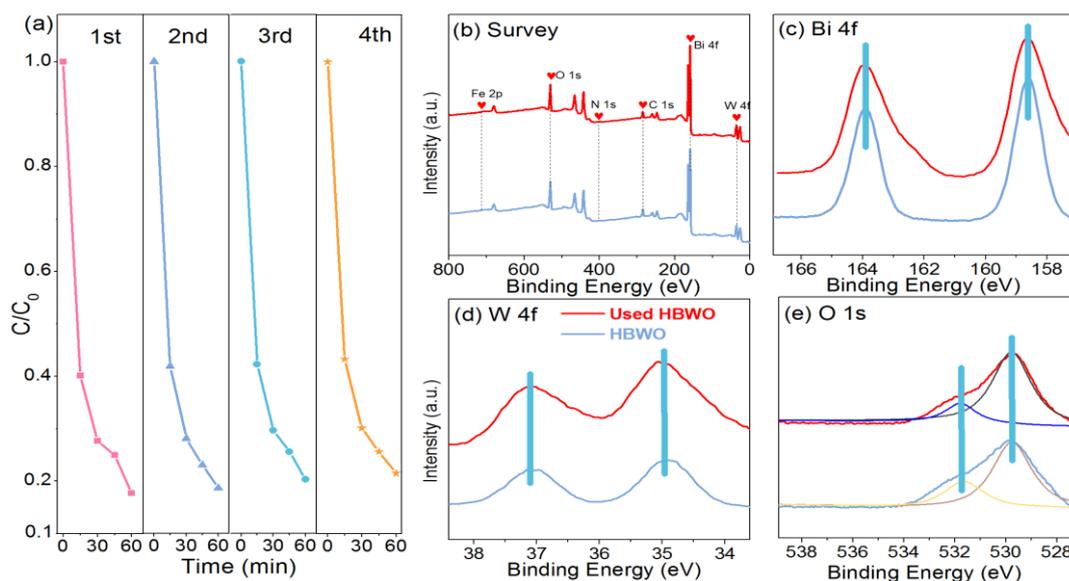


371

372 Fig. 10 Possible photocatalytic reaction mechanism over HBWO composite under solar irradiation.

373 3.3.4 Stability of synthesized catalyst

374 The recyclability and physicochemical stability of HBWO in TC
 375 photodegradation was explored in detail. HBWO was performed four reaction runs,
 376 and the results of the cyclic experiments were shown in Fig. 11a. Obviously, a high
 377 TC photodegradation efficiency of 78.5% can be still obtained over 1HBWO after
 378 four cycles. Besides, XPS spectra of the fresh and used 1HBWO were shown in Fig. 11
 379 (b-e), including the survey spectra, Bi 4f, W 4f and O1s. It was found that the
 380 chemical composites and valence state of 1HBWO kept unchanged after the
 381 photodegradation reaction. Moreover, FTIR and SEM of the used 1HBWO were
 382 detected to further demonstrate the stability (Fig. S4, S5). No apparent changes were
 383 observed. These results affirmed the stability of the synthesized HBWO
 384 photocatalysts in the TC photodegradation process.



385

386 Fig. 11 (a) Cycling runs of TC photodegradation over 1HBWO; XPS analysis of fresh and Used 1HBWO:

387 (b) survey spectra, (c) Bi, (d) W 4f; (e) O 1s.

388 4. Conclusion

389 In summary, HBWO composites with 2D structure were successfully synthesized via a
 390 one-step hydrothermal process. Based on the characterization analysis of the structure,
 391 morphology and photoelectric properties, the composite was formed with hemin
 392 immobilized on the surface of 2D BWO. And the synthesized HBWO showed layered
 393 structure with the interlayer spacing at ~ 0.35 nm. HBWO with mass ratio of
 394 hemin/BWO ranging at 0.5%-5.0% exhibited improved photocatalytic performance for
 395 TC degradation under irradiation. The excellent photocatalytic performance was
 396 attributed to the enhanced photoabsorption and faster electron and oxygen transfer.
 397 According to the ESR analysis and scavenger tests, holes and $\bullet\text{O}_2^-$ played the main
 398 roles in the photocatalytic degradation of TC. It should be noted that the synthesized
 399 HBWO shows high catalytic performance in the environmental pollutant degradation,
 400 which can well meet the requirement of green development.

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